Inactivation of *Ascaris* eggs in water using hydrogen peroxide and a Fenton type nanocatalyst (FeOx/C) synthesized by a novel hybrid production process

Ariadna A. Morales, Rafael Schouwenaars, Heriberto Pfeiffer and Rosa María Ramírez-Zamora

**ABSTRACT**

Inactivation tests of *Ascaris* eggs (Ae) were performed using hydrogen peroxide and a Fenton type nanocatalyst supported on activated carbon (AC) (FeOx/C). Blank inactivation tests were also carried out using H$_2$O$_2$ and H$_2$O$_2$/AC as oxidation systems. The FeOx/C nanocatalyst was synthesized through a novel hybrid method developed in this work. The method is based on the incipient impregnation technique, using isopropyl alcohol as dissolvent and chelating agent of the iron salt and the ultrasonic method. The supported nanocatalyst contained 2.61% w/w of total iron and the support 0.2% w/w. Transmission electron microscopy (TEM)–energy dispersive spectrometer (EDS) images permitted verification of the presence of finely dispersed FeOx nanoparticles, with sizes ranging from 19 to 63 nm. SEM–EDS analysis and TEM images also showed good dispersion of iron oxide nanoparticles, most probably maghemite; $\gamma$-Fe$_2$O$_3$, able to produce hydroperoxyl radical as reported in the literature. The FeOx/C nanocatalyst–H$_2$O$_2$ system showed an average Ae inactivation efficiency of 4.46% Ae/mg H$_2$O$_2$. This value is significantly higher than the result obtained using the support–H$_2$O$_2$ system and H$_2$O$_2$ alone and it is also better than data reported for the classical Fenton process (homogeneous phase) with or without UV light.

**Key words** | helminth, heterogeneous Fenton-like reaction, iron oxides, pathogen, reactive oxygen species

**INTRODUCTION**

Wastewater is a valuable resource for agricultural use in regions where water resources are scarce. However, its reuse may pose substantial health risks for the agriculturists and consumers, particularly in developing countries, due to the presence of pathogens in raw wastewater, such as helminth eggs (Scoot 2008). Navarro & Jiménez (2011) have reported a content of 70–3,000 helminth eggs/L (He/L) in municipal wastewater of developing countries, which is significantly higher than the value recommended by the WHO (1989) for agricultural irrigation (<1 He/L). *Ascaris* eggs (Ae) are the manifestation of the most common helminth species (84%) identified in wastewater and sludge (Jiménez 2007). This parasite affects 25–53% of the world population, mainly in developing countries (Africa, Latin America and the Far East) (Jiménez 2007). Ae are recalcitrant to most disinfection processes due to their shell, consisting of four organic layers of 3–4 $\mu$m thick which protect the eggs, allowing them to remain viable for a long time even under extreme environmental conditions (Brownell & Nelson 2006).

Nelson & Darby (2001) have investigated the validity of using *A. suum*, which infects pigs instead of *A. lumbricoides*, which infects humans, in inactivation tests. Although small genetic differences exist between them, they concluded that their response to inactivation treatments is sufficiently similar to use the more easily

**doi:** 10.2166/wh.2013.203
obtained *A. suum* in large quantities. They also indicate that there is no significant difference in terms of their infectivity and inactivation between eggs dissected from the intestines of mature female worms and eggs isolated from feces.

Advanced oxidation processes (AOPs), especially the Fenton’s reagent, are one of the most efficient technologies for the inactivation of *Ae* (Solís-López 2009; Solís-López et al. 2012). The inactivation of these parasites using the Fenton’s reagent occurs through several reactive oxygen species (ROS), such as superoxide and hydroxyl radicals, produced as a result of the catalytic decomposition of hydrogen peroxide by dissolved iron (II). ROS can degrade the compounds of the protecting layers and consequently reduce the protection of the eggs against aggressive environments. Nonetheless, this process has some disadvantages: it is only efficient in a narrow and highly acid pH range; the catalyst or iron is lost as an acidic sludge with the subsequent economic problems and management of hazardous waste (Garrodo-Ramírez et al. 2010). The heterogeneous Fenton-like process allows these problems to be overcome by using bulk and supported iron species, at micro and nano size scales. This also permits an efficient operation of the Fenton’s reagent at neutral pH (Pignatello et al. 2006; Garrido-Ramírez et al. 2010). It must be mentioned that nanocatalysts are more efficient than micro catalysts, even without using UV light, due to the fact that their catalytic activity improves with the reduction of particle size since the specific contact area is significantly higher than in the case of micro particles (Garrido-Ramírez et al. 2010). The use of supported particles on porous materials permits a higher treatment efficiency of dissolved organic pollutants. This is due to the simultaneous occurrence of efficient oxidation and mechanisms such as adsorption.

Three mechanisms for the catalytic decomposition of H₂O₂ on the surface of iron oxides have been reported in the literature (Lin & Gurol 1998; Andreozzi et al. 2002; Kwan & Voelker 2004). The mechanism proposed by Lin & Gurol (1998) seems to be the most accepted in the academic community. These authors present a complete series of reactions which are also consistent with other heterogeneous Fenton-like reaction mechanisms. The role of the catalyst in this mechanism is to form a surface complex with the hydrogen peroxide molecule. Reversible electron transfer from ligand to metal can then be compensated by the release of the free OH-radical in the case of Fe (II) and the HO₂-radical for Fe (III), named either hydroperoxyl radical or perhydroxyl radical, which is the protonated form of the superoxide radical (*pKₐ* = 4.8).

For the complete sequence of reactions, including termination reactions, the reader is referred to the original work (Lin & Gurol 1998).

Several methods for producing iron oxide nanoparticles supported on porous materials have been reported: precipitation (Adekunle & Ozoemena 2010; Kralj et al. 2010), chemical deposition (Cruat et al. 2011), the thermal decomposition process (González-Arellano et al. 2008; Balu et al. 2010; Wang et al. 2010), the chemical vapor deposition technique (Cordier et al. 2010), sol-gel (Moliner-Martínez et al. 2011), wet impregnation (Lim et al. 2006; Da Silva et al. 2010; Park et al. 2010; Xu et al. 2010; Alam et al. 2011; Bonelli et al. 2011), incipient impregnation (Choi et al. 2007; Quintanilla et al. 2008; Lu et al. 2010; Tsoncheva et al. 2010; Hwang et al. 2011), ion-exchange (Sarkar et al. 2011), and sonochemical process (Jones et al. 2010). Nevertheless, these processes can be very complex, time-consuming, and costly. Thus, investigation into the development of economical production methods is required for the preparation of efficient nanomaterials for use in water treatment.

This work presents the results of *Ae* inactivation tests in water, using hydrogen peroxide and an iron oxide or a Fenton type nanocatalyst supported on activated carbon (AC) (FeOx/C), which was produced by a novel hybrid synthesis method, with the promise of a high cost–benefit ratio.

**METHODS**

The experimental methodology was divided into three main stages: preparation trials of the nanocatalyst, characterization of the material and its support, and inactivation experiments of *Ae* using three oxidation systems (hydrogen peroxide alone, H₂O₂ + support and H₂O₂ + FeOx/C).

**Materials and reagents**

In the synthesis trials, the nanocatalyst’s support was a granular AC of mineral origin (LQ 1000, Carbochem Co.) with
grain sizes ranging from 297 to 590 μm. An alcoholic solution of iron 0.63 M was prepared using Fe(NO₃)₃·9H₂O (99%, Merck) and isopropyl alcohol (grade HPLC, Burdick & Jackson), as iron source and solvent, respectively. The Ti(SO₄)₂ used for the colorimetric method was prepared in our laboratory by acid digestion of TiO₂ (Degussa) and H₂SO₄ (Merck).

For the inactivation tests, hydrogen peroxide at 30% w/w (J. T. Baker) was used. A stock suspension of Ae was prepared in a 0.5% formalin solution with eggs extracted from the uterus of female worms; this suspension, with 92% of initial viability, was stored at 4 °C prior to use. The initial viability of Ae was determined by two techniques, incubation and staining (De Victorica & Galván 2003). Five tests were performed for both, with the results varying 0.8% between the two techniques. A sodium thiosulfate 0.1 N solution was used as received from Hicel (Mexico) for quenching the oxidation reaction.

Preparation of the supported nanocatalyst (FeOx/C)

The FeOx/C nanocatalyst was synthesized by means of a novel hybrid method developed in the present work (Ramírez Zamora et al. 2012). It consists of the incipient impregnation technique, using isopropyl alcohol as dissolvent and chelating agent of the iron salt (Hwang et al. 2011), and the ultrasonic method. For the incipient impregnation process, 1.43 mL of an alcoholic 0.63 M iron solution were added to 1.121 g of AC by dripping from a syringe and mixing continuously at the same time for preparing a 4% w/w supported Fe nanocatalyst (Bonelli et al. 2011). After this, 30 min of ultrasonic treatment (Bransonic 2510R-MT, 100W and 42 kHz ± 6%) was applied for improving the dispersion of the chelated iron salt; this corresponds to the optimal duration as reported by Nagao et al. (2007). Finally, to produce the FeOx nanoparticles, the sample was heated at 85 °C (decomposition temperature of isopropyl alcohol, further referred to as calcination) for 40 min with N₂ as carrier gas (30 cm³/min) in a multifunction system (RIG-100/ISRI). The exhaust gas was analyzed by mass spectroscopy at 8 × 10⁻⁶ torr (Hyden Ltd).

Characterization of materials

The diameter of the iron oxide particles present in AC and FeOx/C nanocatalyst was determined by transmission electron microscopy (TEM) with a JEOL JEM-2010 microscope and the Image® version 1.52 software. The iron oxides in AC and FeOx/C were analyzed by means of X-ray diffraction (XRD). Due to the low XRD detection limit (>3% w/w), it was necessary to produce a second nanocatalyst with 10% w/w Fe (10-FeOx/C) using the same method as described above. AC and 10-FeOx/C were analyzed by XRD using a D5000 (Siemens) diffractometer fitted with a CuKα1 radiation source. The powder patterns were recorded from 2 to 90° with a 0.02° step size. The identification of crystalline phases and minerals was carried out using the database powder diffraction file (PDF-2) of the International Centre for Diffraction Data (ICDD).

The surface area of the materials was determined using N₂ adsorption/desorption isotherms at 77 K with a BelSorp mini II analyzer. Before analysis, samples were heated at 80 °C for 12 h under nitrogen atmosphere for degasification.

The total iron content in AC and FeOx/C was measured by atomic absorption spectroscopy after acid digestion (AAS) using a Spectra AA (Varian) model 220 FS spectrophotometer. Determination of the iron content of AC is important as a reference to determine the amount of Fe incorporated in the support, but also because it may contribute to the hydrogen peroxide decomposition in the heterogeneous Fenton-like reaction, i.e., act as a catalyst by itself.

The morphology and the degree of dispersion of the iron oxides in the support (AC) and FeOx/C nanocatalyst were observed through scanning electron microscopy (SEM) with a Philips XL20 microscope equipped with an EDAX energy dispersive spectrometer (EDS).

Inactivation tests of Ascaris eggs

The catalytic activity of the FeOx/C nanocatalyst was indirectly measured by performing inactivation tests of Ae. Additionally, two inactivation tests of Ae (blanks) were carried out using H₂O₂ alone and H₂O₂/AC as oxidation systems. All the experiments were performed using 500 mL of a helminth egg suspension containing 2 Ae/mL at an initial pH of 4 (Solís-López 2009). The optimal hydrogen peroxide dose (28.64 mg/L) reported for the Fenton reaction in homogeneous phase by Solís-López (2009) and
the Fe: H₂O₂ mass ratio (1:10) reported by Palma et al. (2003) were used in the inactivation tests performed with FeOx/C. The experimental conditions used for AC and FeOx/C were identical. Samples were homogenized in stirred reactors which were protected from light to avoid the photochemical decomposition of hydrogen peroxide. The hydrogen peroxide concentration was measured using the colorimetric method of Eisenberg (1943) at 460 nm with an average deviation of 0.03 mg per 100 mL. The method is based on spectrophotometric measurements of the absorbance of hydrogen peroxide solutions treated with titanium sulfate reagent. The yellow color produced in the reaction is due to the formation of pertitanic acid. After 58 min, the oxidizing action of the hydrogen peroxide residual and ROS was stopped by adding 0.1 mL of sodium thiosulfate in a 0.1 N solution. Then, three samples of 50 mL were collected and filtered on nitrocellulose membranes (8.0 μm pore). The viability of Ae was determined using the vital staining procedure developed by De Victorica & Galván (2003). The viability and inactivation percentages were calculated with the following equations:

\[
\% \text{ Ae viability} = \frac{\text{viable } \text{Ascaris eggs}}{\text{total } \text{Ascaris eggs}} \times 100
\]

\[
\% \text{ of Ae inactivation} = \left( \frac{\% \text{ Ae viability}}{\% \text{ Ae viability}} \right)_{\text{initial}} - \left( \frac{\% \text{ Ae viability}}{\% \text{ Ae viability}} \right)_{\text{final}} \times 100
\]

RESULTS AND DISCUSSION

Preparation of the supported nanocatalyst: iron salt decomposition during nanocatalyst calcination

Quantification of gas emission

The exhaust gas generated during nanocatalyst calcination was used to verify the complete decomposition of nitrate and isopropyl alcohol. Figure 1 shows the mass flow of the exhaust gas measured during the decomposition of the iron salt and isopropyl alcohol (NO, NO₂, CO, CO₂, O₂) in the calcination step of the FeOx/C synthesis. The Fe(NO₃)₃·9H₂O decomposition started while heating the sample. The highest emissions of NO, NO₂, and O₂ were observed after 8 min (54 °C). At this point, a CO₂ signal is observed as well. After 10 min of calcination at 85 °C (boiling point of isopropyl alcohol), a NO and CO₂ signal is also observed and continued until the end of the experiment. Throughout the treatment, no CO emissions were detected, suggesting that isopropyl alcohol combustion was complete. The nitrate decomposition occurred according to the following reaction:

\[
\text{Fe(NO}_3\text{)}_3 \cdot 9\text{H}_2\text{O}_{(s)} \xrightarrow{\Delta} \text{NO}_{(g)} + \text{NO}_2_{(g)} + \text{O}_2_{(g)} + \text{H}_2\text{O}_{(v)} + \text{FeOx}_{(s)}
\]

In accordance with our results, at 85 °C, the most probable iron species present in the nanocatalyst is maghemite (γ-Fe₂O₃). This iron oxide species is not effective for producing hydroxyl radicals (as verified by acid salicylic dosimetry in this work); however, perhydroxyl radical (HO₂⁻) production by maghemite is possible, as described by Lin & Gurol (1998).

Characterization of materials

Iron content – AAS

The iron content of support (AC) and the synthesized nanocatalyst (FeOx/C) was 0.2% w/w and 2.61% w/w,
respectively. The synthesis of the catalyst produces a tenfold increase of the iron content as compared with the support. This value is lower than the theoretical amount (4% w/w) based on the amount of nitrate used, and is similar to the value reported by Bonelli et al. (2011). These authors associate this difference to the precipitation of iron oxides in the flask and syringe used in the impregnation process. Based on this result, the nanocatalyst dose showed the same Fe: H₂O₂ mass ratio (1:10) as reported by Palma et al. (2003) for the Ae inactivation trial.

**Determination of the size and crystalline phase of the iron oxide nanoparticles**

The TEM images of FeOx/C (Figure 2) show irregular iron-containing particles (identified by EDS) with diameters ranging from 19 to 63 nm. These values are slightly larger than the ones reported for nanoparticles synthesized by the synthesis methods for nanocatalysts which most closely resemble the actual one (Quintanilla et al. 2008; Adekunle & Ozoemen 2010; Lu et al. 2010; Wang et al. 2010; Cruzat et al. 2011); the exception is the work of González-Arellano et al. (2008), where larger sizes are reported. The TEM-observations suggest that the supported iron oxides are mainly located on the particle surface.

Not only is the size of the FeOx nanoparticles important, but also their crystalline phase, because the latter plays a determining role in the rate and efficiency of ROS production by the heterogeneous Fenton-like reaction. In the AC sample, only quartz (SiO₂) was identified (JCPDS 046–1045 card). The characteristic reflections of AC between 20–30° and 40–45° (corresponding to the (002) and (001) planes of the graphene stack) were observed. In the 2.1% w/w FeOx/C, it was not possible to detect the diffraction peaks of any iron oxide because its content was below the detection limit of the XRD method. On the other hand, in the 10% w/w Fe catalyst, the XRD pattern suggests the presence of the tetragonal maghemite phase (γ-Fe₂O₃), as evidenced by several small diffraction peaks corresponding to the JCPDS 025–1402 card (Figure 3). This phase has been reported as one of the most active catalysts for the heterogeneous Fenton-like reaction (Wang et al. 2008).

**Figure 2** | TEM micrographs of FeOxC: (a) nanoparticle of 19 nm; (b) nanoparticle of 63 nm.

**Figure 3** | X-ray diffractogram of support (AC) and catalyst 10-FeOxC with SiO₂ (@) and maghemite peaks (@).
Morphology and surface chemical composition of the support

Figure 4 shows a series of SEM images of the AC particles and FeOx/C nanocatalyst. It is noted that a large quantity of bright microparticles are already present in the supporting material AC (Figure 4(a)) which are evidently also seen in the FeOx/C nanocatalyst (Figure 4(c)). These are mainly silicates and aluminosilicates, with a considerable fraction consisting of quartz. Some of the mineral particles contain small fractions of Fe as well. A typical close-up of the surface of the supporting material is presented in Figure 4(b). Similar surfaces can easily be found in the FeOx/C material as well.

Not all the particles in the AC sample are identical and different ‘surface textures’ can be found within the same sample. It is therefore important to compare particles with the same surface appearance when comparing two different samples. The main difference between the starting material

Figure 4 | (a) Supporting material. The white spots in the surface generally are silicate minerals. (b) Detail of the surface of one grain, with a silicate microparticle in the center. (c) FeOx/C. The same white spots as in (a) are seen and are silicate materials. The arrows indicate zones of finely distributed FeOx; it cannot be decided whether these are present on the surface or in the subsurface region by SEM alone. (d) Zone of superficial oxide deposits. (e) Detail of the former; at the amplification given, individual grains cannot be distinguished within the affected zone, indicating that there is either a very fine dispersion of FeOx nanoparticles or a distributed adsorbate of the same material. (f) EDS-spectra of the supporting material (black line), a silicate particle (FeOx/C a) and the iron-rich deposit (FeOx/C b) in (d).
and the supported catalyst is the presence of large, diffuse zones with a brighter tone at the surface and within FeOx/C. A detail of such a zone is shown in Figure 4(d). In Figure 4(e), no individual particles can be observed at high magnifications. The latter image is close to the resolution limit of the SEM, using backscattered electrons. This indicates that FeOx is present either in the form of an adsorbed layer in the carbon or as a fine dispersion of nanoparticles with sizes significantly below the resolution limit of the SEM. On the other hand, TEM results clearly show the presence of the FeOx nanoparticles, the XRD results together with the SEM observations may indicate that at least part of the FeOx is present in the form of an amorphous adsorbate layer on the surface of the support.

**Specific surface area – nitrogen adsorption measurements**

Figure 5 shows the nitrogen adsorption–desorption isotherms of AC and FeOx/C samples, where both exhibited a typical I type isotherm (Leofanti et al. 1998). The adsorption takes place at very low relative pressures because of a strong interaction between pore walls and adsorbate. Although it cannot be excluded that the chemical reaction involved in the synthesis modifies the distribution of the smallest pores, it is more likely that the surface deposition of FeOx as seen in Figures 4(c) and 4(d) closes the access to a significant fraction of the smaller pores.

The average pore diameter was the same (0.2 nm) for both materials. However, the BET surface areas of AC and FeOx/C were different (951.8 and 642.5 m²/g, respectively). It is expected that the decrease of the nanocatalyst’s specific surface area (32.5%) will not negatively influence the production process of ROS because this material shows a significantly (ten times) higher iron content than the support, this latter being the catalyst in the decomposition reaction of hydrogen peroxide. The pore volume of the AC also diminished by 33.7% (from 0.46 cm³/g to 0.31 m³/g). These results indicate that FeOx/C nanoparticles partially block the pores of the support, indicating the predominant location of FeOx/C nanoparticles on the outer surface of the supporting material (Tsoncheva et al. 2010). This textural characteristic of the sample can facilitate and accelerate the primary catalytic decomposition of hydrogen peroxide into the perhydroxyl radical (HO₂•). The results of dosimetric tests, which are not included in this work, using salicylic acid as a dosimeter, did not show the presence of •OH radical.

**Inactivation tests of Ascaris eggs**

Table 1 shows the results of the inactivation tests of Ae, which are compared with data reported in the literature for similar processes. When the FeOx/C-H₂O₂ system is compared with the use of hydrogen peroxide alone, the latter has low inactivation efficiency (0.56% Ae/mg H₂O₂). The support-hydrogen peroxide system without iron oxides showed an even lower catalytic activity (0.084% Ae inactivation/mg H₂O₂) as compared with H₂O₂ alone. This result was expected since iron oxides (catalyst) are virtually absent in the support for the perhydroxyl radical (HO₂•) production; Fe is mostly contained in compact alumino-silicates present in the ash fraction of AC and these silicates have no significant catalytic effect on hydrogen peroxide molecules. In the absence of active iron species for the production of these ROS, an oxidation–reduction reaction between AC carbon (reducing agent) and hydrogen peroxide (oxidant) can take place. In fact, AC is often used in water treatment for the elimination of residual oxidants, such as ozone and chlorine (Gopal et al. 2007).

Several studies have demonstrated an oxidant effect of HO₂•/O₂• radicals on different types of bacteria through a lipid peroxidation mechanism which produces DNA damage (Termini 2000; Rincon & Pulgarin 2006, 2007a,
According to Table 1, the FeOx/C nanocatalyst showed a significantly higher catalytic activity (4.30 ± 0.23% Ae inactivation/mg H₂O₂) than data reported for various iron salts used in the Fenton’s reagent, with similar experimental conditions. The inactivation percentage obtained in this work is between 1.3 and 3.5 times higher than data reported for the homogeneous Fenton-like reaction without UV light (Ramírez et al. 2006a; Solís-López 2009) and it was also more efficient than results obtained with UV light assisted photo-Fenton systems (Ramírez et al. 2006b; Solís-López 2009) and solar light assisted Fenton reactions (Bandala et al. 2011). The reaction time in most of these studies, with the exception of solar Fenton, is similar to the value used in this work. However, the hydrogen peroxide dose in the present work was significantly lower; therefore, the method presented here may present a higher cost–benefit ratio as compared with other Fenton and Fenton-like oxidation systems.

The results of Table 1 indicate that the iron oxide nanoparticles could indeed provide a sufficiently large area of active sites for an efficient H₂O₂ decomposition by the heterogeneous Fenton-like reaction. This can be due to the size and high dispersion of these nanoparticles supported on the AC surface.

An important advantage in the use of the present supported nanocatalyst is that lixiviated iron concentration was very low (0.07 mg/L) and thus no acidic sludge is generated in the process, eliminating the treatment costs of this hazardous waste. Also, this supported nanocatalyst can be recovered by filtration or electromagnetic processes making possible its reuse with a significant cost reduction.

An economic analysis was developed considering the costs for the production of the nanocatalyst and those for the disinfection process (Table 2). For this economic analysis, we considered 4% of inflation rate and 100% of financial leverage with 20% of interest for the nanocatalyst production and 6% of interest for the disinfection process. The disinfection process is self-financed (internal rate of return = zero) in 5 years. Table 2 shows that the cost for wastewater treatment (0.128 USD/m³) is similar to data (0.18 USD/m³) reported by EPA (2006) for a drinking water treatment plant with an average flow of 350 mdg. The operational cost can be reduced by further optimizing the production process of the nanocatalyst and the operation parameters of the disinfection process.

### Table 1 | Percentages of Ae inactivation for homogeneous and heterogeneous Fenton-like reactions

<table>
<thead>
<tr>
<th>Homogeneous Fenton-like reaction</th>
<th>Without UV light</th>
<th>With UV light</th>
<th>Solar</th>
<th>Heterogeneous Fenton-like reaction</th>
<th>Without UV light</th>
</tr>
</thead>
<tbody>
<tr>
<td>3.34</td>
<td>1.01</td>
<td>0.46</td>
<td>1.34</td>
<td>0.18</td>
<td>4.30 ± 0.23</td>
</tr>
<tr>
<td>Total Ae (Ae/mL)</td>
<td>2</td>
<td>2</td>
<td>2</td>
<td>2</td>
<td>0.3</td>
</tr>
<tr>
<td>Doses H₂O₂ (mg/L)</td>
<td>40</td>
<td>28.64</td>
<td>40</td>
<td>21.90</td>
<td>9,520</td>
</tr>
<tr>
<td>Fe/H₂O₂ (w/w)</td>
<td>10</td>
<td>11.24</td>
<td>2</td>
<td>0.07</td>
<td>0.058</td>
</tr>
<tr>
<td>pH</td>
<td>3</td>
<td>3</td>
<td>3</td>
<td>3</td>
<td>3</td>
</tr>
<tr>
<td>Time (min)</td>
<td>50</td>
<td>58.5</td>
<td>50</td>
<td>32.5</td>
<td>120</td>
</tr>
</tbody>
</table>

### Table 2 | Costs of the nanocatalyst production and the disinfection process

<table>
<thead>
<tr>
<th>Nanocatalyst production</th>
<th>Disinfection process</th>
</tr>
</thead>
<tbody>
<tr>
<td>Production</td>
<td>342 ton/year</td>
</tr>
<tr>
<td>Raw materials (USD/day)</td>
<td>$148,089</td>
</tr>
<tr>
<td>Equipment (USD)</td>
<td>$92,683</td>
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<tr>
<td>Wages (USD/year)</td>
<td>$488,049</td>
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<tr>
<td>Operation and management expenses (USD/year)</td>
<td>$10,569</td>
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<tr>
<td>Unit cost</td>
<td>$186,492 USD/ton</td>
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</table>
CONCLUSIONS

A high inactivation rate of Ae (4.46% inactivated Ae/mg H₂O₂) was obtained by using hydrogen peroxide and a Fenton type nanocatalyst. The contribution of the support and H₂O₂ to Ae inactivation was not significant (0.084% Ae inactivation/mg H₂O₂ and 0.56% Ae inactivation/mg H₂O₂, respectively). The FeOx/C inactivation Ae efficiency was higher than other Fenton processes reported in literature up to now.

The nanocatalyst used in the Ae inactivation was prepared through a novel and simple synthesis method developed in this work. This produced well-dispersed iron oxide nanoparticles (most probably maghemite; γ-Fe₂O₃ with sizes of 19–65 nm) supported on AC. Maghemite has been reported as one of the most reactive iron phases for the heterogeneous Fenton-like reaction. The presence of well-dispersed maghemite particles at nanoscale can explain the high catalytic activity of the FeOx/C nanocatalyst synthesized in this work.

ACKNOWLEDGEMENTS

The authors are grateful to PhD Gretchen T. Lapidus Lavine and PhD Gustavo Fuentes Zurita from the Universidad Autónoma Metropolitana – Iztapalapa for the AA analysis and for the use of the RIG-100/ISRI equipment, respectively. They also thank Ivan Puente Lee from the Universidad Nacional Autónoma de Mexico for the TEM analysis. Ariadna A. Morales thanks the CEP-UNAM for the PhD grant. The work was supported by CONACYT under grant No. SEP-2004-C01-48097. The authors thank Gustavo Cadena from the Universidad Nacional Autónoma de Mexico for the economic analysis.

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First received 29 October 2012; accepted in revised form 29 April 2013. Available online 4 June 2013.