

EVALUATION OF IMMUNE FUNCTION IN TWO POPULATIONS OF GREEN SEA TURTLES (*CHELONIA MYDAS*) IN A DEGRADED VERSUS A NONDEGRADED HABITAT

Patricia Sposato,^{1,2} Patricia Keating,¹ Peter L. Lutz,^{1†} and Sarah L. Milton^{1,3}

¹ Department of Biological Sciences, Florida Atlantic University, 777 Glades Rd., Boca Raton, Florida 33431, USA

² Walkabout Ecological Team, Inc., PO Box 690241, Vero Beach, Florida 32969, USA

[†] Deceased

³ Corresponding author (email: smilton@fau.edu)

ABSTRACT: There is a strong correlation between degraded marine habitats and the prevalence of diseases such as green turtle fibropapillomatosis (GTFP) in coastal populations. In GTFP, small to large tumors grow on the turtle's soft tissues and shell, while internal nodules may also occur. The disease primarily affects juvenile green sea turtles (*Chelonia mydas*) that reside in nearshore waters. As a link has been shown between environmental pollution and immune suppression in a variety of animals, the objective of our research was to compare innate and adaptive immune responsiveness in green sea turtles from a severely degraded and a more pristine habitat, which differ greatly in rates of GTFP. We quantified phagocytosis by flow cytometry and performed in vitro stimulation analysis to measure activity of both the innate and adaptive immune systems in wild-caught Florida green turtles. Sea turtles from the degraded environment, both with and without visible cutaneous tumors, exhibited significantly reduced phagocytosis and stimulation indices than did those from the less polluted environment. Our results suggest that environmental factors may contribute to the development of GTFP and thus can impact the health of sea turtle populations.

Key words: Adaptive immune function, flow cytometry, green turtle fibropapillomatosis (GTFP), innate immune function, lymphocyte proliferation, phagocytosis.

INTRODUCTION

Sea turtles are bound to coastal habitats during several developmental stages, thus the integrity of the nearshore ecosystem may play a critical role in their welfare. Pollutants in nearshore habitats, including carcinogens and heavy metals (Perrault et al. 2011; Sinaei and Bolouki 2017; Speer et al. 2018), polychlorinated biphenyls (Camacho et al. 2013), harmful algal blooms (Perrault et al. 2014, 2017a), and agricultural runoff (Van Houtan et al. 2010; Bossart 2011) may threaten sea turtle health. These anthropogenic factors are linked to infectious diseases in coastal wildlife (Schaefer et al. 2011; Vilela et al. 2016; Bossart et al. 2019). In marine mammals, these include tumors and disease in beluga whales (*Delphinapterus leucas*; Lair et al. 2016; Iqbal et al. 2018) and deaths among California sea lions (*Zalophus californianus*) from toxic algal blooms (Lefebvre et al. 2016, 2018) and infectious disease (Seguel et al. 2019). In sea turtles, a high incidence of green

turtle fibropapillomatosis (GTFP) is associated with degraded habitats (Herbst and Klein 1995; dos Santos et al. 2010). Investigators have therefore examined the associations between GTFP and various environmental factors including the prevalence of tumor-promoting toxins (Landsberg et al. 1999; Arthur et al. 2008), persistent organic pollutants (Keller et al. 2014; Sanchez-Sarmiento et al. 2017), dietary changes (Van Houtan et al. 2014), and ultraviolet radiation (Duffy et al. 2018), in addition to the presence of herpesvirus (Lackovich et al. 1999; Rodenbusch et al. 2014).

While considered the most likely etiologic agent for GTFP, chelonid alpha herpesvirus 5 has coexisted with sea turtle populations for 300 million yr, with no evidence for a recent increase in virulence correlating to the increased disease prevalence over the past century (Herbst et al. 2004; Lawrance et al. 2018). The virus is present worldwide (Greenblatt et al. 2005) in all hard-shelled sea turtle species (Quackenbush et al. 1998; Alfaro-

Núñez et al. 2014; Chaves et al. 2017) and in clinically healthy turtles (Page-Karjian et al. 2012; Alfaro-Núñez and Gilbert 2014). This suggests that the virus is widespread, but can be latent or subclinical (Alfaro-Núñez et al. 2016), and that expression of the disease is multi-factorial, involving interactions between the virus, host, and environment (Jones et al. 2016; Duffy and Martindale 2019). One factor that may affect host-virus interactions is a decrease in immune competence linked to pollutants, as has been shown in many animals including shellfish (Jiang et al. 2017), fish (Martyniuk et al. 2016; Chen et al. 2019), freshwater turtles (Ming-ch'eng Adams et al. 2016), and marine mammals (reviewed in Desforgues et al. 2016). Cray et al. (2001) found reduced immune responsiveness in captive green sea turtles (*Chelonia mydas*) exhibiting GTFP compared to animals without evident tumors, though some have suggested that immune suppression is a result of, rather than a contributing agent to, the disease (Work et al. 2001).

We examined immune function in two populations of noncaptive turtles, comparing resident turtles from an area of poor water quality with those in a more pristine environment. Florida's Indian River Lagoon (IRL) is a heavily polluted estuary with high levels of heavy metals and persistent organic pollutants (Wang et al. 1992; Durden et al. 2007; Fair et al. 2010) and eutrophication (Lapointe et al. 2015; Barile 2018). As with other animals in the IRL, resident juvenile green turtles exhibit high rates of disease, with approximately 50% of green turtles showing tumors (Hirama and Ehrhart 2007; Lawrance et al. 2018). However, animals from the more pristine Trident Basin (TRI), located near Cape Canaveral, Florida, have essentially no GTFP (Hirama and Ehrhart 2007). While most previous papers on immune function in sea turtles examined adaptive immunity (Cray et al. 2001; Work et al. 2001; Keller et al. 2014; Rousselet et al. 2017), innate immune function is also likely to be important (Rousselet et al. 2013); thus, we examined aspects of both innate and adaptive function in the two populations. Adaptive immunity involves the specific rec-

ognition of antigens and the development of memory cells and, in turtles, is most often measured by lymphocyte proliferation (Rousselet et al. 2013), while innate immunity acts as an initial defense mechanism against pathogenic agents and involves natural killer cells and phagocytic cells. We hypothesized that cells from both the innate and adaptive immune system would show reduced function in animals from the more polluted habitat.

MATERIALS AND METHODS

Collections sites

The IRL is a 250-km long shallow estuary on the east-central coast of Florida, with minimal water exchange with the Atlantic Ocean. The TRI is a man-made embayment located near the mouth of an inlet with a strong tidal flux, is located on government property, and experiences little pollution. The TRI is considered a representative location for a pristine, unimpacted population and is free of GTFP (Hirama and Ehrhart 2007). Because turtles captured in the IRL may or may not have internal nodules, and thus gross observation cannot determine if they are GTFP positive, for the purposes of this study we categorized animals captured in the IRL by the presence or absence of visible tumors (VT+ and VT-, respectively) without presumption of viral infection.

Animal capture and handling

Blood samples were obtained from 87 green turtles captured at TRI ($n=27$) and in the IRL ($n=60$) as part of ongoing tag-recapture studies at those locations (Florida Fish and Wildlife Conservation Commission permit no. FWC MTP186 and National Marine Fisheries Service permit no. NMFS 14506). Animals were sampled in 1999 and 2001 ($n=52$) and from 2011 to 2013 ($n=35$) across all seasons. Animals were caught in a 152-m tangle net, dip netted, and removed to a boat for data collection. Animals were released following collection of biologic and morphometric data, blood collection, and tagging. All immunologic analyses were performed under Florida Fish and Wildlife Conservation Commission permit no. FWC MTP053 and approved by the Florida Atlantic University Institutional Animal Care and Use Committee. For each study, 2–5 mL of blood were drawn from the dorsal cervical sinus and transferred to a sodium-heparin Vacutainer (BD Biosciences, Franklin Lakes, New Jersey, USA). Blood samples were chilled above ice (approximately 4 C) for transportation to Florida Atlantic University for analysis.

Hematology

Heparinized whole blood smears were stained using Diff Quick differential stain (Sigma-Aldrich, St. Louis, Missouri, USA) or a 1:20 diluted Giemsa stain (Sigma-Aldrich). Leukocytes were counted under 40 \times magnification and recorded as a percentage of the total leukocyte population. Additionally, an aliquot of 10 μ L of whole blood was centrifuged at 27,950 \times G for 5 min to determine packed cell volume (PCV).

Separation of whole blood

Whole blood was layered on a discontinuous Percoll gradient (GE, Pittsburgh, Pennsylvania, USA): a 60% layer to restrict monocytes and lymphocytes (peripheral blood mononuclear cells; PBMCs) and a 75% layer to separate granulocytes (polymorphonuclear cells; PMNs). Each layer was washed twice in 1 \times phosphate-buffered saline. The viable cell yield was determined by standard trypan blue hemocytometry. Each sample was then diluted with phosphate-buffered saline to bring the concentration to 1 \times 10⁶ cells/mL.

Phagocytosis assay

Samples were incubated with a suspension of 1.0 μ m fluorescein isothiocyanate (FITC)-labeled latex beads (Spherotech, Lake Forest, Illinois, USA). Then 50- μ L of FITC-labelled beads were added to cells (100 μ L PBMCs and PMNs) in Hank's balanced salt solution (Fisher Scientific, Waltham, Massachusetts, USA) in 96-well plates. After 1 h, phagocytosis was interrupted by placing samples on ice. Leukocytes were then gently layered on cold 10% bovine serum albumin (VWR, Radnor, Pennsylvania, USA) and centrifuged at 800 \times G at 4 C for 8 min to elute noninternalized and nonspecifically bound beads. The percent of cells containing phagocytosed beads was evaluated using a FACSCalibur flow cytometer (BD Biosciences) and analyzed with FlowJo software (BD Biosciences). Fluorescence was measured at 480 nm. Cells alone, without fluorescent beads, were used as a negative control. Forward scatter vs. side scatter plots were used to identify PMN, PBMC, and red blood cell populations by size and granularity (Rousselet et al. 2013). For each sample, 20,000 events were collected.

In vitro lymphocyte proliferation

Lymphocytes were isolated from whole blood with Histopaque-1077 (Sigma-Aldrich) and washed three times. The viable cell yield was determined and each sample diluted with complete media to 1 \times 10⁶/mL. Lymphocytes were incubated at 37 C and 5% CO₂ with phytohe-

magglutinin or concanavalin A—considered to be T-cell mitogens; lipopolysaccharide (LPS; a B-cell mitogen); phorbol 12-myristate 13-acetate (PMA) + ionomycin; or pokeweed (PWM) mitogen (both a T- and B-cell mitogen), with complete media as a negative control (Cray et al. 2001). All stimulating agents were from Sigma-Aldrich. After 48 h, cells were pulsed with 20 μ L of Promega CellTiter96[®] Aqueous One Solution Cell Proliferation Assay (Promega, Madison, Wisconsin). Optical densities were then taken for each plate utilizing the Wallac 1250 Betaplate liquid scintillation counter (Perkin Elmer, Waltham, Massachusetts, USA) every hour for 3 h. Stimulation indices were calculated using counts per minute (cpm) as [mean mitogen cpm]/[mean medium only cpm].

Statistical analysis

Statistical comparisons were made both between sites, time of year (season), and VT+ vs. VT- status. Data were tested for normality and equal variance of residuals, with appropriate analyses for parametric (Shapiro-Wilkes) or non-parametric (Kruskal-Wallis and Dunn's) data to detect differences between groups. Statistics were run on Sigmaplot software version 12.1 (IBM SPSS Statistics for Windows, Armonk, New York, USA).

RESULTS

Hematology

Leukocyte differential and PCV were determined for each sample. In both IRL and TRI groups, heterophils comprised the largest fraction of the leukocyte population; lymphocytes made up the next largest fraction followed by eosinophils. Monocytes made up the smallest fraction (Fig. 1). Interestingly, the percentage of circulating monocytes was significantly higher in VT+ (IRL) turtles ($P=0.002$) than TRI turtles. There were no other significant differences in the relative proportions of circulating leukocytes among the populations ($P=0.119$), nor was mean PCV significantly different between TRI and IRL turtles either with or without tumors.

Phagocytosis assays

Flow cytometry revealed that both the PBMC and PMN fractions of sea turtle white blood cells are capable of phagocytosis (Fig.

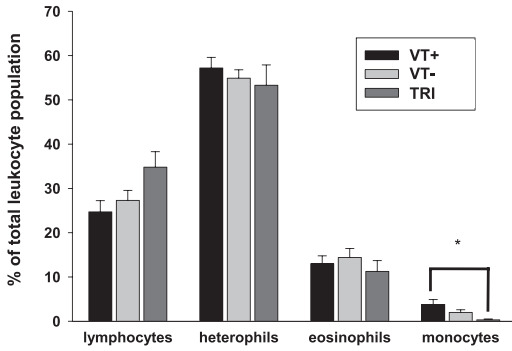


FIGURE 1. Predominant leukocyte populations for green turtles (*Chelonia mydas*) turtles from two locations in Florida, USA: the Indian River Lagoon (high prevalence of green turtle fibropapillomatosis [GTFP]; with [VT+] and without [VT-] visible tumors) and the Trident Basin (TRI) where GTFP is rare. Animals were sampled in 1999 and 2001 and from 2011 to 2013 across all seasons. There were no differences in the percentage composition of the leukocytes between turtles from the two locations ($P=0.866$), except in the monocyte population. *Monocyte levels are significantly higher in TRI turtles than in VT+ ($P=0.002$). There is no statistical difference ($P=0.119$) between VT+ turtles and turtles without tumors (VT- and TRI).

2). The mean percentage of phagocytosis for PBMCs and PMNs across all groups was 4% and 8.2%, respectively. Overall, immune function as determined by phagocytosis was

significantly related to geographic location, tumor status, and season.

Phagocytosis was significantly higher for both PBMC and PMN populations from TRI turtles than for leukocytes from VT+ or VT- animals captured in the IRL ($P<0.05$; Fig. 3). Regarding tumor status, the PBMCs in turtles with GTFP (consisting of animals from both the TRI population and IRL VT-) exhibited a higher percentage of phagocytosis than did those animals with visible tumors ($P<0.001$). Within the IRL population, PBMC activity was significantly higher in VT- animals than in VT+ turtles ($P<0.05$); PMN activity in the IRL population also appeared to be higher in VT- compared to VT+ animals, but the difference was not statistically significant (Fig. 3).

Additionally, there was a seasonal difference in phagocytic capacity: for leukocyte populations in IRL turtles (both VT+ and VT-), there was a higher fraction of phagocytically active cells in the summer samples (June–August) than in winter (January–March; Fig. 4). In winter, samples from TRI turtles had significantly higher percentages of phagocytosis (mean 10.5%) compared to both VT- (3.5%, $P<0.05$) and VT+ (0.7%, $P<0.001$) turtles from the IRL. The TRI

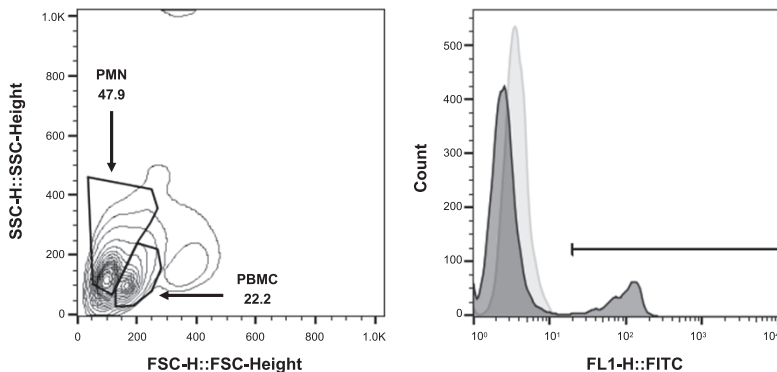


FIGURE 2. Representative analysis of flow cytometry results for phagocytes of green turtles (*Chelonia mydas*) sampled in the Indian River Lagoon, Florida, USA, where there is a high prevalence of green turtle fibropapillomatosis. Animals were sampled from 2011 to 2013 across all seasons. Contour plot (left panel) shows forward scatter (FSC) vs. side scatter (SSC), with areas containing peripheral blood mononuclear cells (PBMCs) and polymorphonuclear cells (PMNs) outlined. Histogram (right panel) shows each cell population peak and representative gate for cells that have phagocytosed fluorescein isothiocyanate-labelled (FITC+) beads. The darker gray represents the PMNs and the lighter gray represents the PBMCs. Each sample represents 20,000 events.

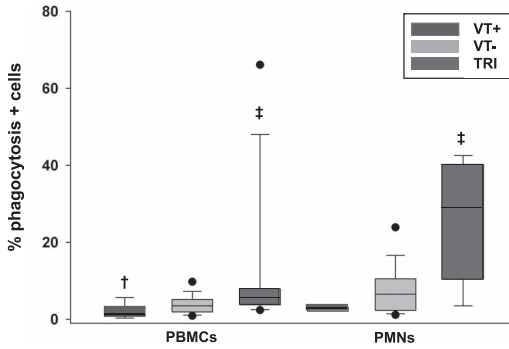


FIGURE 3. Box and whisker plots showing percentage of phagocytosis-positive cells in the peripheral blood mononuclear cell (PBMC) and polymorphonuclear cell (PMN) layers from all green turtles (*Chelonia mydas*) captured in the Indian River Lagoon (IRL; $n=60$) and Trident Basin (TRI; $n=27$), Florida, USA. Animals were sampled from 2011 to 2013 across all seasons. Leukocytes of TRI turtles had significantly greater phagocytosis by both the PBMC and PMN subpopulations than either IRL group ($P<0.05$). Within the IRL, the PBMCs had greater phagocytosis in turtles without visible tumors (VT-) than did those animals with visible tumors (VT+; $P<0.001$). †=Within same white blood cell (WBC) population, significantly different from both VT+ and TRI turtles. ‡=Within same WBC population, significantly different from both VT+ and VT- turtles.

population was not sampled in the summer months. Additionally, VT- turtles in the IRL exhibited significantly higher percentages of phagocytic cells than did VT+ turtles during both the summer and winter months ($P=0.023$); VT+ turtles exhibited the lowest percentages of phagocytosis regardless of season ($P<0.001$).

In vitro lymphocyte proliferation

Sea turtle leukocytes responded most strongly to PMA + ionomycin, with a lesser response to either LPS or PWM (Fig. 5). The only mitogen resulting in a statistically significant difference between TRI turtles and both VT- and VT+ turtles in the IRL was PMA + ionomycin ($P=0.045$). Additionally, VT+ turtles had stimulation indices that were significantly lower than those of nonpapilloma turtles from either the IRL or TRI when pulsed with LPS or PWM ($P<0.05$). The median stimulation index in response to phytohemagglutinin for leukocytes from VT+

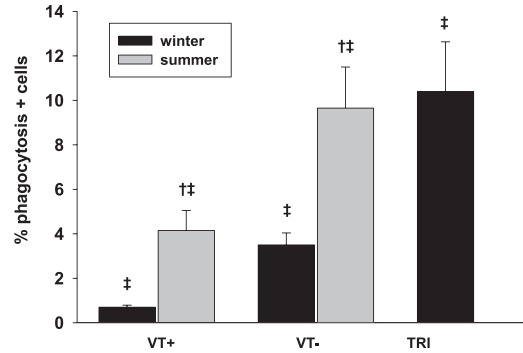


FIGURE 4. Mean percentages of phagocytosis in leukocytes collected in winter vs. summer in green (*Chelonia mydas*) turtles of the Indian River Lagoon (IRL, Florida). Animals were sampled from 2011 to 2013. There were no summer data collected from the Trident Basin (TRI, Florida), but winter rates of phagocytosis are significantly higher ($P<0.001$) than in animals from the IRL. Additionally, IRL turtles without visible tumors (VT-) exhibited higher rates of phagocytosis than did those with visible tumors (VT+) during summer months ($P=0.023$). †=Significantly different between groups in same season. ‡=Significantly different between seasons for same group.

turtles was lower than in TRI animals, but the difference was not significant. Overall responsiveness was low in response to concanavalin A, and there was no difference between turtle

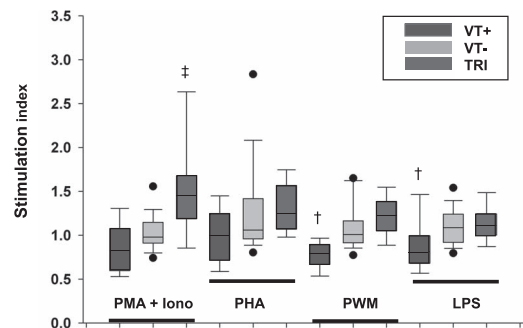


FIGURE 5. Box and whisker plots showing stimulation indices for green sea turtle (*Chelonia mydas*) lymphocytes exposed to respective mitogens. ‡=Significantly different from Indian River Lagoon (IRL, Florida) turtles both with (VT+) and without visible tumors (VT-). †=Significantly different from both Trident Basin (TRI, Florida) and IRL VT-. Stimulating agents included phytohemagglutinin (PHA), concanavalin A, lipopolysaccharide (LPS), phorbol 12-myristate 13-acetate (PMA) + ionomycin, or pokeweed mitogen (PWM), with complete media as a negative control.

populations (data not shown). When IRL VT+ and VT- turtles were grouped to focus on location rather than tumor status, IRL animals on the whole did not respond as vigorously to either PMA + ionomycin or PWM as did those animals from TRI ($P < 0.045$).

DISCUSSION

Our results indicate that both adaptive and innate immune function are compromised in green turtles captured in the highly polluted IRL, where historical rates of GTFP in juvenile green turtles are $\geq 50\%$ (Hirama and Ehrhart 2007; Lawrance et al. 2018). Within the IRL group, immune function was generally lower in VT+ turtles than in VT- turtles. By comparison, turtles from the TRI basin are free of GTFP, and both the innate and adaptive branches of immune function exhibit greater immune competence. This correlation suggests that location or habitat quality may contribute to immune competence in green turtles; thus, disease prevalence may reflect ecosystem health. While Work et al. (2001) suggested that immune suppression is a result rather than a contributing agent to the disease, our data suggest that habitat quality, disease state, and immune function are intertwined, forming a positive feedback loop wherein polluted environments impact the immune system and make animals more prone to the expression of GTFP, which in turn further compromises the immune system. Links between pollution and disease have been well described in mammalian studies (Scholin et al. 2000; Lefebvre et al. 2016, 2018; Lair et al. 2016; Iqbal et al. 2018; Seguel et al. 2019). In Florida, for example, an overabundance of nitrogen in the IRL encourages the proliferation of the fungus *Paracoccidioides brasiliensis* (Durden et al. 2009) and increasing numbers of Atlantic bottlenose dolphins (*Tursiops truncatus*) with chronic mycotic dermal infection (Vilela et al. 2016; Bossart et al. 2019), while pollutants in Charleston Harbor, South Carolina, have been linked to the spread of infectious disease in their resident dolphin population (Bossart et al. 2017; Reif et al. 2017).

Similar to resident dolphin populations, juvenile green turtles are intrinsically bound to coastal habitats and subject to a variety of ecologic stressors that influence the health of both the ecosystem and its inhabitants. This is especially true in the IRL, an estuary experiencing declining health due to multiple ecologic stressors. The IRL experiences very little tidal mixing with the Atlantic Ocean; freshwater inputs and nutrient pollution have exacerbated harmful algal blooms (HABs; Gobler and Sunda 2012; O'Neil et al. 2012), including a brown algae bloom in 2011 that resulted in the loss of approximately 60% of total seagrass cover (St. Johns River Water Management District 2012) and a blue-green algal bloom in 2019 that increased toxic microcystin levels. Such HAB events are associated with declining health in resident *Tursiops* populations (Twiner et al. 2011). Other organisms and estuaries are also affected: decreased salinity in the IRL has also resulted in, for example, an overabundance of a toxic fungus, *Aphanomyces invadens*, which creates lesions in the skin and muscle tissue of fish (Vandersea et al. 2006; Sosa et al. 2007), while novel neoplasias have recently been detected in fish in other Florida estuaries (Kiryu et al. 2018). Prevalence of GTFP is generally associated with degraded habitats and nearshore environments worldwide (Milton and Lutz 2003; dos Santos et al. 2010), and juvenile green turtles in the IRL have shown an increase in GTFP prevalence during long-term monitoring (Aguirre and Lutz 2004). An examination of stress responses at the molecular level also suggests that green turtles in the IRL are physiologically stressed. Whether they have visible tumors or not, levels of cellular stress markers are higher in these animals than in TRI turtles (Deming 2008).

Hematology

Studies in other aquatic organisms have shown impacts by pollutants such as HABs and heavy metals on immune function, including in manatees (*Trichechus manatus*; Walsh et al. 2015), freshwater turtles (Walsh

et al. 2019), and sea turtles (Walsh et al. 2010; Perrault et al. 2014, 2017a, 2017b), while differences in circulating leukocytes in response to pollution have been demonstrated in fish (Marchand et al. 2020). One aspect of immune competence that might reflect ecosystem health is a difference in circulating leukocyte populations, and previous studies have shown variable heterophil to lymphocyte ratios based on sampling site (Aguirre et al. 1994; Cray et al. 2001; Lutz et al. 2001). We found that heterophils comprised the majority of white blood cells, as reported in other sea turtle studies (Rousselet et al. 2013; Muñoz et al. 2014; Rossi et al. 2016); interestingly, we observed an increase in the mean heterophil population from 34% (in 1999) to 53% (in 2014) in the TRI turtles, possibly indicating some other physiologic stress. In reptiles, heterophils and monocytes are thought to be the first line of immune defense (Rousselet et al. 2013), although a subset of B-cells that normally produce antibodies has also been shown to have phagocytic capabilities in freshwater turtles (Zimmerman et al. 2009). Monocytes are also an important component of the innate immune response and comprise up to 5% of the leukocyte population in IRL turtles with visible tumors, 2% in IRL turtles without apparent tumors, and 0.8% in turtles from the TRI, though in one study of captive loggerheads, monocyte levels were greater than eosinophil numbers (Rousselet et al. 2013).

While the small fraction of monocytes in the leukocyte population compared to the heterophil population would suggest that monocytes are not the primary phagocytic cell in juvenile green sea turtles, Rousselet et al. (2013) reported that monocytes had the highest phagocytic activity in captive juvenile loggerhead turtles, while Rossi et al. (2016) found that in green turtles with GTFP, the lymphocyte and monocyte populations had significantly more phagocytic activity than did the heterophils. Differences between studies are likely to result from a variety of factors including use of wild caught vs. captive turtles, the length of residence in rehabilitation facilities, the selection of phagocytic

target and opsonization, and whether other stimulants were used (e.g., Zymosan A from yeast [Rossi et al. 2016]). Although overall monocyte numbers were low in this study, there was a significant difference between IRL animals with tumors and the TRI turtles, suggesting that this white blood cell subpopulation is upregulated in cases of active GTFP. Increased monocyte numbers in reptiles are thought to indicate chronic conditions such as bacterial or parasitic infections, inflammation, and neoplastic diseases (Stacy et al. 2011).

Phagocytosis

Flow cytometry revealed that both monocytes and heterophils were capable of ingesting FITC beads. Phagocytosis by different fractions of the white blood cell populations, including heterophils, monocytes, and eosinophils, has been reported previously (Rossi et al. 2009, 2016; Rousselet et al. 2013; Muñoz et al. 2014). Rousselet et al. (2013) reported similar levels of phagocytosis for both monocytes and heterophils in juvenile loggerhead turtles. Though Rossi et al. (2016) reported in one study that lymphocytes and monocytes had greater phagocytic activity than did granulocytes in green turtles, method differences may have resulted in different ratios of leukocytes: the heterophil percentage in the Rossi et al. (2016) study (10%) was much lower than the lymphocyte population (88.3%), suggesting that few heterophils were retrieved. In our study, phagocytic activity varied dependent on where the animals were captured, tumor status, and time of year.

While there was no significant difference in the circulating heterophil percentages between IRL and TRI turtles, heterophils (the main component of the PMN layer in our study) from TRI exhibited the highest percentages of phagocytosis, indicating that immune cells in those turtles from a more pristine environment perform better than those from a degraded habitat. Even within the IRL population, VT- turtles showed better immune competence than did those with visible tumors, supporting the hypothesis

that diseased states may hamper immune function, as suggested by Work et al. (2001).

Within the IRL, tumor-bearing animals are more prevalent in the summer months (Hirama and Ehrhart 2007) despite the fact that, as shown in this study, phagocytosis is significantly higher in summer than winter, suggesting better immune function in the warmer months. Many factors could result in higher apparent rates of GTFP in the summer IRL population, from a different group of resident turtles to a higher growth rate of some unknown disease vector such as leeches (Herbst and Klein 1995). Among green turtles from the IRL, VT⁻ animals did exhibit significantly higher percentages of phagocytosis in both seasons than did VT⁺ turtles (Fig. 4).

In vitro lymphocyte proliferation

The lymphocyte proliferation assays also showed a significant difference in adaptive immune function between the IRL and TRI turtles. Cells from TRI turtles exhibited the highest stimulation indices with PMA + ionomycin and PWM, both of which are reflective of a response from both B- and T-cells, though because we did not separately test the B- and T-cells in this study, there was no way to differentiate their activity. Lymphocytes from IRL turtles exhibited distinctly reduced lymphocyte proliferation in response to mitogen stimulation compared to TRI turtles. For the majority of mitogens, the only significant difference was between VT⁺ turtles in the IRL and TRI animals; the IRL VT⁻ animals consistently showed lower responses than TRI animals but were higher than VT⁺ turtles, suggesting that both T and B lymphocytes from the IRL turtles with visible tumors were relatively unresponsive. The lack of significant differences may be due in part to the overall low responsiveness of turtle leukocytes to the mitogens, as was seen in earlier studies (Lutz 2001; Cray et al. 2001). The reagent that generated the strongest response in all populations (PMA + ionomycin) was also the one that showed a significant difference not only between TRI and VT⁺

animals, but also between TRI and VT⁻, and between VT⁺ and VT⁻ turtles in the IRL. Previous studies of adaptive immune responses in sea turtles have shown correlations to suppressed lymphocyte proliferation in turtles exposed to organochlorides, disease (GTFP), and degraded environments (Cray et al. 2001; Lutz et al. 2001; Keller et al. 2006). Negative effects on immune function by various contaminants in sea turtles have been demonstrated specifically for mercury (Day et al. 2007), other heavy metals (Camacho et al. 2013), and polychlorinated biphenyls (Rousselet et al. 2017), while both papilloma disease and herpes viral infections are associated with reduced immune competence in animals (Nicholls and Stanley 2000), from oysters (de Lorgeril et al. 2018) to dogs (Sundberg et al. 1994) and cows (Jones 2019).

Together, the results of the phagocytosis and lymphocyte proliferation assays substantiate the hypothesis that turtles from more pristine habitats are better able to mount a robust defense against pathogens, while those from degraded habitats exhibit reduced immune function that may make them more prone to disease. Defects in innate immune function are likely to increase susceptibility to pathogenic infection, as host leukocytes would be unable to clear microbial cells and dead or damaged host cells. Additionally, stimulation of the adaptive immune system by the innate response may be reduced (Abbas et al. 2012). This study supports the earlier findings of Cray et al. (2001) that an altered adaptive immune function is associated with tumor development and, in fact, strongly suggests that immune responsiveness may be linked to environmental health, not just tumor status. We also found that VT⁻ turtles from either the TRI or IRL site showed better immune competence than those with visible FP tumors, even within the highly polluted IRL, agreeing with Work et al. (2001) that reduced immune function follows disease. Our results, however, also indicate that reduced immune competence may initially permit disease, and disease status in turn may then further hinder immunocompetence. Such a vicious cycle could explain why certain locations have such

a high incidence of disease, while other areas have clinically healthy turtles that test positive for chelonid alphaherpesvirus 5 (Page-Karjian et al. 2012; Alfaro-Núñez and Gilbert 2014). These results are not surprising because immunosuppression is known to stimulate further tumor growth and increase infection risk (Schreiber et al. 2011). Our study indicates that where increased incidence of disease exists, gross observation alone does not necessarily indicate that wildlife populations are healthy, which in turn may reflect overall ecosystem health.

ACKNOWLEDGMENTS

This research was conducted under contract to FAU by the National Marine Fisheries Service, Southwest Fisheries Science Center, Honolulu Laboratory, Hawaii (1999), and with grants from the Morris Animal Foundation and Friends of Gumbo Limbo Nature Center, Boca Raton, Florida (2012). The authors thank Carolyn Cray from the University of Miami Miller School of Medicine for her invaluable contributions and review of the earlier manuscript.

LITERATURE CITED

- Abbas AK, Lichtman AH, Pillai S. 2012. *Cellular and molecular immunology*. 7th Ed. Elsevier Saunders, Philadelphia, Pennsylvania, 525 pp.
- Aguirre A, Balazs GH, Zimmerman B, Galey FD. 1994. Organic contaminants and trace metals in the tissues of green turtles (*Chelonia mydas*) afflicted with fibropapillomas in the Hawaiian islands. *Mar Pollut Bull* 28:109–114.
- Aguirre AA, Lutz P. 2004. Marine turtles as sentinels of ecosystem health: Is fibropapillomatosis an indicator? *EcoHealth* 1:275–283.
- Alfaro-Núñez A, Bojesen AM, Bertelsen MF, Wales N, Balazs GH, Gilbert MTP. 2016. Further evidence of Chelonid herpesvirus 5 (ChHV5) latency: High levels of ChHV5 DNA detected in clinically healthy marine turtles. *PeerJ* 4:e2274.
- Alfaro-Núñez A, Frost Bertelsen M, Bojesen AM, Rasmussen I, Zepeda-Mendoza L, Tange Olsen M, Gilbert MTP. 2014. Global distribution of Chelonid fibropapilloma-associated herpesvirus among clinically healthy sea turtles. *BMC Evol Biol* 14:206.
- Alfaro-Núñez A, Gilbert MTP. 2014. Validation of a sensitive PCR assay for the detection of Chelonid fibropapilloma-associated herpesvirus in latent turtle infections. *J Virol Methods* 206:38–41.
- Arthur K, Limpus C, Balazs G, Capper A, Udy J, Shaw G, Keuper-Bennett U, Bennett P. 2008. The exposure of green turtles (*Chelonia mydas*) to tumour promoting compounds produced by the cyanobacterium *Lyngbya majuscula* and their potential role in the aetiology of fibropapillomatosis. *Harmful Algae* 7:114–125.
- Barile PJ. 2018. Widespread sewage pollution of the Indian River Lagoon system, Florida (USA) resolved by spatial analyses of macroalgal biogeochemistry. *Mar Pollut Bull* 128:557–574.
- Bossart G, Fair P, Schaefer A, Reif J. 2017. Health and Environmental Risk Assessment Project for bottlenose dolphins *Tursiops truncatus* from the southeastern USA. I. Infectious diseases. *Dis Aquat Organ* 125:141–153.
- Bossart GD. 2011. Marine mammals as sentinel species for oceans and human health. *Vet Pathol* 48:676–690.
- Bossart GD, Romano TA, Peden-Adams MM, Schaefer AM, Rice CD, Fair PA, Reif JS. 2019. Comparative innate and adaptive immune responses in Atlantic bottlenose dolphins (*Tursiops truncatus*) with viral, bacterial, and fungal infections. *Front Immunol* 10:1125.
- Camacho M, Luzardo OP, Boada LD, López Jurado LF, Medina M, Zumbado M, Orós J. 2013. Potential adverse health effects of persistent organic pollutants on sea turtles: Evidences from a cross-sectional study on Cape Verde loggerhead sea turtles. *Sci Total Environ* 458–460:283–289.
- Chaves A, Aguirre AA, Blanco-Peña K, Moreira-Soto A, Monge O, Torres AM, Soto-Rivas JL, Lu Y, Chacón D, Fonseca L, et al. 2017. Examining the role of transmission of chelonid alphaherpesvirus 5. *EcoHealth* 14:530–541.
- Chen J, Xu Y, Han Q, Yao Y, Xing H, Teng X. 2019. Immunosuppression, oxidative stress, and glycometabolism disorder caused by cadmium in common carp (*Cyprinus carpio* L.): Application of transcriptome analysis in risk assessment of environmental contaminant cadmium. *J Hazard Mater* 366:386–394.
- Cray C, Varella R, Bossart GD, Lutz PL. 2001. Altered in vitro immune responses in green turtles (*Chelonia mydas*) with fibropapillomatosis. *J Zoo Wildl Med* 32:436–440.
- Day RD, Segars AL, Arendt MD, Lee AM, Peden-Adams MM. 2007. Relationship of blood mercury levels to health parameters in the loggerhead sea turtle (*Caretta caretta*). *Environ Health Perspect* 115:1421–1428.
- de Lorigeril J, Lucasson A, Petton B, Toulza E, Montagnani C, Clerissi C, Vidal-Dupiol J, Chaparro C, Galinier R, Escoubas J-M, et al. 2018. Immunosuppression by OsHV-1 viral infection causes fatal bacteraemia in Pacific oysters. *Nat Commun* 9:4215.
- Deming AC. (2008). *Stress protein and anti-apoptotic protein expression in green turtle (Chelonia mydas) fibropapillomatosis*. Master's Thesis, Biological Sciences, Florida Atlantic University, Boca Raton, Florida, 78 pp.
- Desforges J-PW, Sonne C, Levin M, Siebert U, De Guise S, Dietz R. 2016. Immunotoxic effects of environmental pollutants in marine mammals. *Environ Int* 86:126–139.

- dos Santos R, Martins A, Torezani E, Baptistotte C, Farias J, Horta P, Work T, Balazs G. 2010. Relationship between fibropapillomatosis and environmental quality: A case study with *Chelonia mydas* off Brazil. *Dis Aquat Organ* 89:87–95.
- Duffy DJ, Martindale MQ. 2019. Perspectives on the expansion of human precision oncology and genomic approaches to sea turtle fibropapillomatosis. *Commun Biol* 2:54.
- Duffy DJ, Schnitzler C, Karpinski L, Thomas R, Whilde J, Eastman C, Yang C, Krstic A, Rollinson D, Zirkelbach B, et al. 2018. Sea turtle fibropapilloma tumors share genomic drivers and therapeutic vulnerabilities with human cancers. *Commun Biol* 1:63.
- Durden WN, St. Leger J, Stolen M, Mazza T, Londono C. 2009. Lacaziosis in bottlenose dolphins (*Tursiops truncatus*) in the Indian River Lagoon, Florida, USA. *J Wildl Dis* 45:849–856.
- Durden WN, Stolen MK, Adams DH, Stolen ED. 2007. Mercury and selenium concentrations in stranded bottlenose dolphins from the Indian River Lagoon system, Florida. *Bull Mar Sci* 81:37–54.
- Fair PA, Adams J, Mitchum G, Hulsey TC, Reif JS, Houde M, Muir D, Wirth E, Wetzel D, Zolman E, et al. 2010. Contaminant blubber burdens in Atlantic bottlenose dolphins (*Tursiops truncatus*) from two southeastern US estuarine areas: Concentrations and patterns of PCBs, pesticides, PBDEs, PFCs, and PAHs. *Sci Total Environ* 408:1577–1597.
- Gobler CJ, Sunda WG. 2012. Ecosystem disruptive algal blooms of the brown tide species, *Aureococcus anophagefferens* and *Aureoumbra lagunensis*. *Harmful Algae* 14:36–45.
- Greenblatt RJ, Quackenbush SL, Casey RN, Rovnak J, Balazs GH, Work TM, Casey JW, Sutton CA. 2005. Genomic variation of the fibropapilloma-associated marine turtle herpesvirus across seven geographic areas and three host species. *J Virol* 79:1125–1132.
- Herbst L, Ene A, Su M, Desalle R, Lenz J. 2004. Tumor outbreaks in marine turtles are not due to recent herpesvirus mutations. *Curr Biol* 14:R697–R699.
- Herbst LH, Klein PA. 1995. Green turtle fibropapillomatosis: Challenges to assessing the role of environmental cofactors. *Environ Health Perspect* 103 (4 Suppl): 27–30.
- Hirama S, Ehrhart LM. 2007. Description, prevalence, and severity of green turtle fibropapillomatosis in three developmental habitats on the east coast of Florida. *Florida Sci* 70:435–448.
- Iqbal A, Measures L, Lair S, Dixon B. 2018. *Toxoplasma gondii* infection in stranded St. Lawrence Estuary beluga *Delphinapterus leucas* in Quebec, Canada. *Dis Aquat Organ* 130:165–175.
- Jiang Y, Tang X, Sun T, Wang Y. 2017. BDE-47 exposure changed the immune function of haemocytes in *Mytilus edulis*: An explanation based on ROS-mediated pathway. *Aquat Toxicol* 182:58–66.
- Jones C. 2019. Bovine herpesvirus 1 counteracts immune responses and immune-surveillance to enhance pathogenesis and virus transmission. *Front Immunol* 10:1008.
- Jones K, Ariel E, Burgess G, Read M. 2016. A review of fibropapillomatosis in green turtles (*Chelonia mydas*). *Vet J* 212:48–57.
- Keller JM, Balazs GM, Nilsen F, Rice M, Work TM, Jensen BA. 2014. Investigating the potential role of persistent organic pollutants in Hawaiian green sea turtle fibropapillomatosis. *Environ Sci Technol* 48: 7807–7816.
- Keller JM, McClellan-Green PD, Kucklick JR, Keil DE, Peden-Adams MM. 2006. Effects of organochlorine contaminants on loggerhead sea turtle immunity: Comparison of a correlative field study and in vitro exposure experiments. *Environ Health Perspect* 114: 70–76.
- Kiryu Y, Landsberg J, Bakenhaster M, Tyler-Jedlund A, Wilson P. 2018. Putative histiocytic sarcoma in redbfin needlefish *Strongylura notata* (Belontiiformes: Belontiidae) in Florida, USA. *Dis Aquat Organ* 132:57–78.
- Lackovich JK, Brown DR, Homer BL, Garber RL, Mader DR, Moretti RH, Patterson AD, Herbst LH, Oros J, Jacobson ER, et al. 1999. Association of herpesvirus with fibropapillomatosis of the green turtle *Chelonia mydas* and the loggerhead turtle *Caretta caretta* in Florida. *Dis Aquat Organ* 37:89–97.
- Lair S, Measures LN, Martineau D. 2016. Pathologic findings and trends in mortality in the beluga (*Delphinapterus leucas*) population of the St. Lawrence Estuary, Quebec, Canada, from 1983 to 2012. *Vet Pathol* 53:22–36.
- Landsberg JH, Balazs GH, Steidinger KA, Baden DG, Work TM, Russell DJ. 1999. The potential role of natural tumor promoters in marine turtle fibropapillomatosis. *J Aquat Anim Health* 11:199–210.
- Lapointe BE, Herren LW, Debortoli DD, Vogel MA. 2015. Evidence of sewage-driven eutrophication and harmful algal blooms in Florida's Indian River Lagoon. *Harmful Algae* 43:82–102.
- Lawrance MF, Mansfield KL, Sutton E, Savage AE. 2018. Molecular evolution of fibropapilloma-associated herpesviruses infecting juvenile green and loggerhead sea turtles. *Virology* 521:190–197.
- Lefebvre KA, Hendrix A, Halaska B, Duignan P, Shum S, Isoherranen N, Marcinek DJ, Gulland FMD. 2018. Domoic acid in California sea lion fetal fluids indicates continuous exposure to a neuroteratogen poses risks to mammals. *Harmful Algae* 79:53–57.
- Lefebvre KA, Quakenbush L, Frame E, Huntington KB, Sheffield G, Stimmelmayer R, Bryan A, Kendrick P, Ziel H, Goldstein T, et al. 2016. Prevalence of algal toxins in Alaskan marine mammals foraging in a changing arctic and subarctic environment. *Harmful Algae* 55:13–24.
- Lutz PL, Cray C, Sposato P. 2001. Studies of the association between immunosuppression and fibropapillomatosis within three habitats of *Chelonia mydas*. *NOAA Technical Memorandum Administrative Report H-01-01C*. National Marine Fisheries

- Service, Scientific Publications Office, Seattle, Washington, 26 pp.
- Marchand A, Tebby C, Beaudouin R, Catteau A, Porcher JM, Turiès C, Bado-Nilles A. 2020. Reliability evaluation of biomarker reference ranges for mesocosm and field conditions: Cellular innate immunomarkers in *Gasterosteus aculeatus*. *Sci Total Environ* 698:134333.
- Martyniuk CJ, Doperalski NJ, Prucha MS, Zhang J-L, Kroll KJ, Conrow R, Barber DS, Denslow ND. 2016. High contaminant loads in Lake Apopka's riparian wetland disrupt gene networks involved in reproduction and immune function in largemouth bass. *Comp Biochem Physiol Part D Genomics Proteomics* 19: 140–150.
- Milton SL, Lutz PL. 2003. Physiological and genetic responses to environmental stress. In: *The biology of sea turtles*, Vol. 2, Lutz PL, Musick JA, Wyneken J, editors. CRC Press, Boca Raton, Florida, pp. 163–198.
- Ming-ch'eng Adams CI, Baker JE, Kjellerup BV. 2016. Toxicological effects of polychlorinated biphenyls (PCBs) on freshwater turtles in the United States. *Chemosphere* 154:148–154.
- Muñoz FA, Franco-Noguez SY, Gonzalez-Ballesteros E, Negrete-Philippe AC, Flores-Romo L. 2014. Characterisation of the green turtle's leukocyte subpopulations by flow cytometry and evaluation of their phagocytic activity. *Vet Res Commun* 38:123–128.
- Nicholls PK, Stanley MA. 2000. The immunology of animal papillomaviruses. *Vet Immunol Immunopathol* 73:101–127.
- O'Neil JM, Davis TW, Burford MA, Gobler CJ. 2012. The rise of harmful cyanobacteria blooms: The potential roles of eutrophication and climate change. *Harmful Algae* 14:313–334.
- Page-Karjian A, Torres F, Zhang J, Rivera S, Diez C, Moore PA, Moore D, Brown C. 2012. Presence of chelonid fibropapilloma-associated herpesvirus in tumored and non-tumored green turtles, as detected by polymerase chain reaction, in endemic and non-endemic aggregations, Puerto Rico. *Springerplus* 1:35.
- Perrault J, Wyneken J, Thompson LJ, Johnson C, Miller DL. 2011. Why are hatching and emergence success low? Mercury and selenium concentrations in nesting leatherback sea turtles (*Dermochelys coriacea*) and their young in Florida. *Mar Pollut Bull* 62:1671–1682.
- Perrault JR, Schmid JR, Walsh CJ, Yordy JE, Tucker AD. 2014. Brevetoxin exposure, superoxide dismutase activity and plasma protein electrophoretic profiles in wild-caught Kemp's ridley sea turtles (*Lepidochelys kempii*) in southwest Florida. *Harmful Algae* 37:194–202.
- Perrault JR, Stacy NI, Lehner AF, Mott CR, Hirsch S, Gorham JC, Buchweitz JP, Bresette MJ, Walsh CJ. 2017a. Potential effects of brevetoxins and toxic elements on various health variables in Kemp's ridley (*Lepidochelys kempii*) and green (*Chelonia mydas*) sea turtles after a red tide bloom event. *Sci Total Environ* 605–606:967–979.
- Perrault JR, Stacy NI, Lehner AF, Poor SK, Buchweitz JP, Walsh CJ. 2017b. Toxic elements and associations with hematology, plasma biochemistry, and protein electrophoresis in nesting loggerhead sea turtles (*Caretta caretta*) from Casey Key, Florida. *Environ Pollut* 231:1398–1411.
- Quackenbush SL, Work TM, Balazs GH, Casey RN, Rovnak J, Chaves A, duToit L, Baines JD, Parrish CR, Bowser PR, et al. 1998. Three closely related herpesviruses are associated with fibropapillomatosis in marine turtles. *Virology* 246:392–399.
- Reif J, Schaefer A, Bossart G, Fair P. 2017. Health and Environmental Risk Assessment Project for bottlenose dolphins *Tursiops truncatus* from the southeastern USA. II. Environmental aspects. *Dis Aquat Organ* 125:155–166.
- Rodenbusch CR, Baptistotte C, Werneck MR, Pires TT, Melo MTD, De Ataíde MW, Dos Reis KDHL, Testa P, Alieue MM, Canal CW. 2014. Fibropapillomatosis in green turtles *Chelonia mydas* in Brazil: Characteristics of tumors and virus. *Dis Aquat Organ* 111: 207–217.
- Rossi S, de Queiroz Hazarbasanov NGT, Sánchez-Sarmiento AM, Prioste FES, Matushima ER. 2016. Immune response of green sea turtles with and without fibropapillomatosis: Evaluating oxidative burst and phagocytosis via flow cytometry. *Chelonian Conserv Biol* 15:273–278.
- Rossi S, Sá-Rocha V, Kinoshita D, Genoy-Puerto A, Zwarg T, Werneck M, Sá-Rocha L, Matushima E. 2009. Flow cytometry as a tool in the evaluation of blood leukocyte function in *Chelonia mydas* (Linnaeus, 1758) (Testudines, Cheloniidae). *Brazilian J Biol* 69: 899–905.
- Rousselet E, Levin M, Gebhard E, Higgins BM, DeGuise S, Godard-Codding CAJ. 2013. Evaluation of immune functions in captive immature loggerhead sea turtles (*Caretta caretta*). *Vet Immunol Immunopathol* 156:43–53.
- Rousselet E, Levin M, Gebhard E, Higgins BM, DeGuise S, Godard-Codding CAJ. 2017. Polychlorinated biphenyls (PCBs) modulate both phagocytosis and NK cell activity *in vitro* in juvenile loggerhead sea turtles (*Caretta caretta*). *J Toxicol Environ Health A* 80:556–561.
- Sanchez-Sarmiento AM, Rossi S, Vilca FZ, Thijl Vanstreels RE, Monteiro SH, Vale LAS, Dos Santos RG, Marigo J, Bertozzi CP, Grisi Filho JHH, et al. 2017. Organochlorine pesticides in green sea turtles (*Chelonia mydas*) with and without fibropapillomatosis caught at three feeding areas off Brazil. *J Mar Biol Assoc UK* 97:215–223.
- Schaefer AM, Bossart GD, Mazzoil M, Fair PA, Reif JS. 2011. Risk factors for colonization of *E. coli* in Atlantic bottlenose dolphins (*Tursiops truncatus*) in the Indian River Lagoon, Florida. *J Environ Public Health* 2011:597073.
- Scholin CA, Gulland F, Doucette GJ, Benson S, Busman M, Chavez FP, Cordaro J, DeLong R, De Vogelaere A, Harvey J, et al. 2000. Mortality of sea lions along

- the central California coast linked to a toxic diatom bloom. *Nature* 403:80–84.
- Schreiber RD, Old LJ, Smyth MJ. 2011. Cancer immunoediting: Integrating immunity's roles in cancer suppression and promotion. *Science* 331:1565–1570.
- Seguel M, Colegrove KM, Field C, Whoriskey S, Norris T, Duignan P. 2019. Polyphasic rhabdomyositis in California sea lions (*Zalophus californianus*): Pathology and potential causes. *Vet Pathol* 56:619–629.
- Sinaei M, Bolouki M. 2017. Metals in blood and eggs of green sea turtles (*Chelonia mydas*) from nesting colonies of the northern coast of the Sea of Oman. *Arch Environ Contam Toxicol* 73:552–561.
- Sosa ER, Landsberg JH, Stephenson CM, Forstchen AB, Vandersea MW, Litaker RW. 2007. *Aphanomyces invadans* and ulcerative mycosis in estuarine and freshwater fish in Florida. *J Aquat Anim Health* 19:14–26.
- Speer RM, Wise CF, Young JL, Aboueissa A-M, Martin Bras M, Barandiaran M, Bermúdez E, Márquez-D'Acunti L, Wise JP. 2018. The cytotoxicity and genotoxicity of particulate and soluble hexavalent chromium in leatherback sea turtle lung cells. *Aquat Toxicol* 198:149–157.
- Stacy NI, Alleman AR, Sayler KA. 2011. Diagnostic hematology of reptiles. *Clin Lab Med* 31:87–108.
- St. John's River Water Management District. 2012. *Indian River Lagoon 2012 superbloom plan of investigation*. <http://floridaswater.com/itsyourlagoon/initiative.html>. Accessed June 2019.
- Sundberg JP, Smith EK, Herron AJ, Jenson AB, Burk RD, Van Ranst M. 1994. Involvement of canine oral papillomavirus in generalized oral and cutaneous verrucosis in a Chinese Shar Pei dog. *Vet Pathol* 31:183–187.
- Twiner MJ, Fire S, Schwacke L, Davidson L, Wang Z, Morton S, Roth S, Balmer B, Rowles TK, Wells RS. 2011. Concurrent exposure of bottlenose dolphins (*Tursiops truncatus*) to multiple algal toxins in Sarasota Bay, Florida, USA. *PLoS One* 6:e17394.
- Van Houtan KS, Hargrove SK, Balazs GH. 2010. Land use, macroalgae, and a tumor-forming disease in marine turtles. *PLoS One* 5:e12900.
- Van Houtan KS, Smith CM, Dailer ML, Kawachi M. 2014. Eutrophication and the dietary promotion of sea turtle tumors. *PeerJ* 2:e602.
- Vandersea MW, Litaker RW, Yonnish B, Sosa E, Landsberg JH, Pullinger C, Moon-Butzin P, Green J, Morris JA, Kator H, et al. 2006. Molecular assays for detecting *Aphanomyces invadans* in ulcerative mycotic fish lesions. *Appl Environ Microbiol* 72:1551–1557.
- Vilela R, Bossart GD, St. Leger JA, Dalton LM, Reif JS, Schaefer AM, McCarthy PJ, Fair PA, Mendoza L. 2016. Cutaneous granulomas in dolphins caused by novel uncultivated *Paracoccidioides brasiliensis*. *Emerg Infect Dis* 22:2063–2069.
- Walsh CJ, Butawan M, Yordy J, Ball R, Flewelling L, de Wit M, Bonde RK. 2015. Sublethal red tide toxin exposure in free-ranging manatees (*Trichechus manatus*) affects the immune system through reduced lymphocyte proliferation responses, inflammation, and oxidative stress. *Aquat Toxicol* 161:73–84.
- Walsh CJ, Cocilova C, Restivo J, Flewelling L, Milton S. 2019. Immune function in *Trachemys scripta* following exposure to a predominant brevetoxin congener, PbTx-3, as a model for potential health impacts for sea turtles naturally exposed to brevetoxins. *Ecotoxicology* 28:1085–1104.
- Walsh CJ, Leggett SR, Carter BJ, Colle C. 2010. Effects of brevetoxin exposure on the immune system of loggerhead sea turtles. *Aquat Toxicol* 97:293–303.
- Wang T, Hoffman M, David J, Parkinson R. 1992. Chlorinated pesticide residue occurrence and distribution in mosquito control impoundments along the Florida Indian River Lagoon. *Bull Environ Contam Toxicol* 49:217–223.
- Work TM, Rameyer RA, Balazs GH, Cray C, Chang SP. 2001. Immune status of free-ranging green turtles with fibropapillomatosis from Hawaii. *J Wildl Dis* 37:574–581.
- Zimmerman LM, Vogel LA, Edwards KA, Bowden RM. 2009. Phagocytic B cells in a reptile. *Biol Lett* 6:270–273.

Submitted for publication 17 November 2020.

Accepted 8 March 2021.