

Modeling caffeine adsorption by multi-walled carbon nanotubes using multiple polynomial regression with interaction effects

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ABSTRACT

Permanent monitoring of environmental issues demands efficient, accurate, and user-friendly pollutant prediction methods, particularly from operating variables. In this research, the efficiency of multiple polynomial regression in predicting the adsorption capacity of caffeine (q) from an experimental batch mode by multi-walled carbon nanotubes (MWCNTs) was investigated. The MWCNTs were specified by scanning electron microscope, Fourier transform infrared spectroscopy and point of zero charge. The results confirmed that the MWCNTs have a high capacity to uptake caffeine from the wastewater. Five parameters including pH, reaction time (t), adsorbent mass (M), temperature (T) and initial pollutant concentration (C) were selected as input model data and q as the output. The results indicated that multiple polynomial regression which employed C , M and t was the best model (normalized root mean square error = 0.0916 and $R^2 = 0.996$). The sensitivity analysis indicated that the predicted q is more sensitive to the C , followed by M , and t . The results indicated that the pH and temperature have no significant effect on the adsorption capacity of caffeine in batch mode experiments. The results displayed that estimations are slightly overestimated. This study demonstrated that the multiple polynomial regression could be an accurate and faster alternative to available difficult and time-consuming models for q prediction.

Key words | adsorption capacity, multiple polynomial regression, sensitivity analysis

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INTRODUCTION

Caffeine (CFN) with chemical formula $C_8H_{10}N_4O_2$ is an alkaloid pertaining to the methylxanthine group naturally available in foods including tea, coffee, kola nuts and cacao beans. In humans, caffeine acts not only as a central nervous system provocative but also as a natural pesticide for the plant (Pavel *et al.* 2003). Caffeine is liberated in the aquatic environment and has been found in surface water, ground water and also in wastewater effluents in large concentration ($\sim 10 \text{ g L}^{-1}$) (Glassmeyer *et al.* 2005). Because caffeine is difficult to metabolize and is commonly found in ground water and surface water, it has been proposed as a chemical indicator of environmental pollution. For

instance, zebra fish embryos could not survive when the dose of caffeine in water was greater than 300 mg L^{-1} (Chen *et al.* 2008). Literature also indicated that this substance is toxic to most of the aquatic environment (Pollack *et al.* 2009). Caffeine can damage soil fertility, cause barriers to seed germination and lowers growth of seedlings (Pollack *et al.* 2009). Thus, elimination of excess caffeine from different water sources is necessary.

Different chemical and physical technologies have been employed for the elimination of this substance in water including coagulation/flocculation, ion exchange, oxidation/reduction, membrane separation, biological

treatment and adsorption (Ma et al. 2012). Traditional procedures applied for caffeine elimination are either expensive or involve the utilization of toxic organic materials. However, as traditional wastewater treatment do not degrade caffeine, it is important to look for alternatives (Álvarez-Torrellas et al. 2016). Adsorption procedure is one of the best wastewater treatments due to the simplicity, ease of operation, high efficiency and can be utilized in small scale household units; thus, adsorption techniques are commonly applied. Adsorbents must have suitable characteristics including high adsorption capacity, surface area, selectivity, lifetime and capacity for regeneration (Arshadi et al. 2014).

Nowadays, a great deal of interest is being concentrated on the use of nano-structured materials such as carbon nanotubes (CNTs) as adsorbents to eliminate harmful and toxic organic pollutant from aqueous media. CNTs which were described by Iijima (1991), are one of the most commonly investigated carbon nano-materials and can act as good adsorbents for the removal of toxic pollutants due to their layered and hollow structure, and also large surface area (Tan et al. 2012). CNTs materials can be divided into three groups: functionalized CNTs, single-wall CNTs and multi-wall CNTs (MWCNTs) (Yu et al. 2014). MWCNTs are successfully applied in the removal of many pollutants such as methyl red (Ghaedi & Kokhdan 2012), methyl orange (Hosseini et al. 2011), Eriodrome Cyanine R (Ghaedi et al. 2011), pharmaceuticals and personal care products (Wang et al. 2016) from wastewater. However, examples of the utilization of MWCNTs for caffeine removal are rather scarce in the literature.

The main factors affecting the adsorption reaction including pH, reaction time (t), adsorbent mass (M), temperature (T) and initial pollutant concentration (C) were investigated and analyzed, using the equations governing the adsorption process such as adsorption isotherm or/and adsorption kinetic models. In these equations only one factor is variable and the others are held constant. It is better to use an equation that encompasses all factors affecting the adsorption process. In other words, it is better to investigate the effects of all factors simultaneously. Several researchers have employed artificial intelligence models to predict adsorption efficiency of pollutant from aqueous media. These models include artificial neural networks

(Yetilmezsoy & Demirel 2008; Gamze Turan et al. 2011a, 2011b; Amiri et al. 2013a), adaptive neural-based fuzzy inference systems (Amiri et al. 2013b), wavelet neural networks and support vector regression (Mousavi et al. 2012). However, many cases of prediction, such as adsorption process polynomial regression, did at least as well if not better than the artificial intelligence models. The main purpose of this research was data modeling of caffeine removal from aqueous media using MWCNTs by multiple polynomial regression with interaction effects.

MATERIALS AND METHODS

Adsorbate and adsorbent

Caffeine (i.e 1,3,7-trimethylpurine-2,6-dione) was purchased from Sigma–Aldrich Co. (Germany), in analytical purity and applied in the experiments directly without any further purification. Suitable concentrations of caffeine solutions were provided by diluting a stock solution with deionized water. MWCNTs were prepared by Nanocyl Co. (Belgium) and used as adsorbent. All other chemicals were purchased from Merck (Germany). Some of the physical characteristics of MWCNTs are seen in Table 1.

Characterization techniques

The pH of solution was modified using 0.1 M HCl/NaOH using a pH meter (Metrohm, 827 pH Lab). Zero point charge (pH_{ZPC}) of MWCNTs was measured with the solid addition method (Balistreri & Murray 1981). Nitrogen (99.999%) adsorption tests were carried out at $-196\text{ }^{\circ}\text{C}$

Table 1 | Physical properties of MWCNTs

Items	MWCNTs
Diameter	10–20 nm
Length	30 μm
Purity	>95 wt%
Ash	<1.5 wt%
Specific surface area ($\text{m}^2\text{ g}^{-1}$)	200
Density (g cm^3)	2.1

using volumetric apparatus (Quantachrome NOVA automated gas sorption analyzer). The concentrations of caffeine solutions were obtained by using a UNICO-2100 UV-Vis spectrophotometer at a wavelength corresponding to the maximum absorbance, $\lambda_{max} = 275$ nm. The functional groups existing in MWCNTs was investigated using the Fourier transform infrared (FTIR spectroscopy) technique. FTIR spectra were recorded using a Jasco FT/IR-680 plus spectrophotometer as KBr pellets. The size and surface morphology of MWCNTs were determined using a scanning electron microscope (SEM) (MIRA3TESCAN-XMU).

Experimental procedure

Laboratory batch experiments were carried out to study the adsorption of caffeine on MWCNTs. The experiments were performed at different temperatures (25–80 °C) using several bottles under various experimental variables including pH (2–12), reaction time (1–30 min), the adsorbent mass (0.02–0.15 g), and the initial caffeine concentration (0.5–400 mg L⁻¹). The bottles containing the mixture of caffeine solution with MWCNTs were shaken at 120 rpm using a SEBD001 rotary orbital shaker. In the final stage of batch experiments, the solutions were filtered using filter paper No. 42. The concentration of residual caffeine in the remaining solution was determined using a UV/Vis spectrophotometer. The adsorption capacity (q) (mg g⁻¹) of caffeine using MWCNTs was computed using Equation (1),

$$q = \frac{C_0 - C_e}{m} \times V \quad (1)$$

where C_0 and C_e are the initial concentration of caffeine in solution and the concentration in equilibrium (mg L⁻¹), respectively, V is the volume of solution (L) and m is the adsorbent mass (g).

Multiple polynomial regression with interaction effects

Statistical procedures such as multiple polynomial regression are the best tools for studying any relationship between low example sizes of response and independent variables (Razi & Athappilly 2005). Statistical analysis that

justifies the combinations of factor levels was used through multiple polynomial regression analysis using Minitab 17 software. Multiple polynomial regression is a procedure applied to model the relationship between one or more independent parameters and a response variable. To create this regression model, it was necessary to determine critical parameters, given that the best model must have the least number of necessary parameters and maximum accuracy. In this research, q was the response variable. Also t , C , M , pH and T were continuous predictors. To implement the multiple polynomial regression, the measured data set included 114 caffeine samples which were divided into two parts: the first group (86 observations, 75% of the data) was randomly selected for building the model, and the second group (28 observations, 25% of the data) was applied for testing the model.

Goodness of fitted model

In this work, the goodness of multiple polynomial regression was determined by the linear coefficient of determination (R^2), normalized root mean square error (NRMSE) and mean residual error (MRE). These indices are described as follows (Willmott et al. 1985):

$$R^2 = \frac{[\sum_{i=1}^n (q_i - \bar{q})(\hat{q}_i - \bar{\hat{q}})]^2}{\sum_{i=1}^n (q_i - \bar{q})^2 \sum_{i=1}^n (\hat{q}_i - \bar{\hat{q}})^2} \quad (2)$$

$$NRMSE = \left[\frac{\sum_{i=1}^n (\hat{q}_i - q_i)^2}{n (\bar{q})^2} \right]^{0.5} \quad (3)$$

$$MRE = \frac{\sum_{i=1}^n (\hat{q}_i - q_i)}{n \bar{q}} \quad (4)$$

where q_i and \hat{q}_i are the measured and model estimated amounts of caffeine adsorbed (mg g⁻¹), respectively, \bar{q} and $\bar{\hat{q}}$ are the mean of measured and predicted values, respectively, and n is the number of measurements. A lower $NRMSE$ value and higher R^2 value are regarded as showing goodness of agreement between measured and estimated caffeine adsorbed data. $NRMSE$ of less than 10% represents accuracy of the model. When the quantity of $NRMSE$ is

closer to zero the precision of results is higher. The prediction is considered poor if the NRMSE is higher than 30%, fair if the NRMSE is higher than 20% and lower than 30%, good if the NRMSE is higher than 10% and lower than 20% and excellent if the NRMSE is lower than 10% (Jamieson *et al.* 1991). MRE is a measure of estimation bias, with positive and negative MRE values demonstrating overestimation and underestimation, respectively.

RESULTS AND DISCUSSION

Characterization of MWCNTs

The surface and textural morphology of MWCNTs by SEM image is illustrated in Figure 1. As shown in Figure 1, MWCNTs size are in the range of 10–20 nm in diameter and $>30\ \mu\text{m}$ in length. The well developed porous structure can be observed. Figure 1(a) shows quite clearly MWCNTs, but after adsorption (B), the surface is tarnished, suggesting that a thin layer of caffeine covers MWCNTs, without changing the morphology of MWCNTs.

Figure 2 shows the FTIR spectra of the adsorbent before and after caffeine adsorption. The carbonyl band at $1,637.4\ \text{cm}^{-1}$ is the most intense band. This band is related to the presence of C=C in the aromatic rings of caffeine molecules and corresponds to the stretching vibration of C=O bonds.

This peak shows caffeine adsorption on MWCNTs according to Figure 3. The broad band at $3,435\ \text{cm}^{-1}$ is

characteristic of alcohol O-H stretch and amine N-H stretch groups present in the adsorbent surface. In the FTIR spectra of the MWCNTs, there is a small peak at $2,927.8\ \text{cm}^{-1}$ assigned to the carboxylic acid O-H stretch and alkyl C-H stretch groups. These peaks show minor chemical reactions between the caffeine molecules and MWCNTs surfaces. The peak at $1,402.18\ \text{cm}^{-1}$ is characteristic of aromatic C-H bending. The physical properties of the MWCNTs show that the surface area and mean pore diameter are $200\ \text{m}^2\ \text{g}^{-1}$ and 3.46 nm, respectively. However, the great specific surface area and average pore size of MWCNT give an increase in the approachability of active sites of the adsorbent in contact with contaminated water.

In order to better comprehend the net charge of the adsorbent surface at the different pH, the pH_{ZPC} of the MWCNTs were measured. The results of pH_{ZPC} determination are shown in Figure 4. pH_{ZPC} of the MWCNTs was estimated to be about 3.82 which indicates that the surface charge of the adsorbent is negative at high pH of solution. The pH of the solution determines not only the predominant species in the solution but also the net charge on the carbonaceous materials. This aspect has been studied by Ayranci *et al.* (2005) on phthalic acid and its esters adsorption onto activated carbon cloth.

Several works have been performed in order to elucidate the mechanism of adsorption of many molecules on different adsorbents. Those publications reveal that adsorption of organic molecules from dilute aqueous solutions on carbon materials is a complex interplay between electrostatic and non-electrostatic interactions

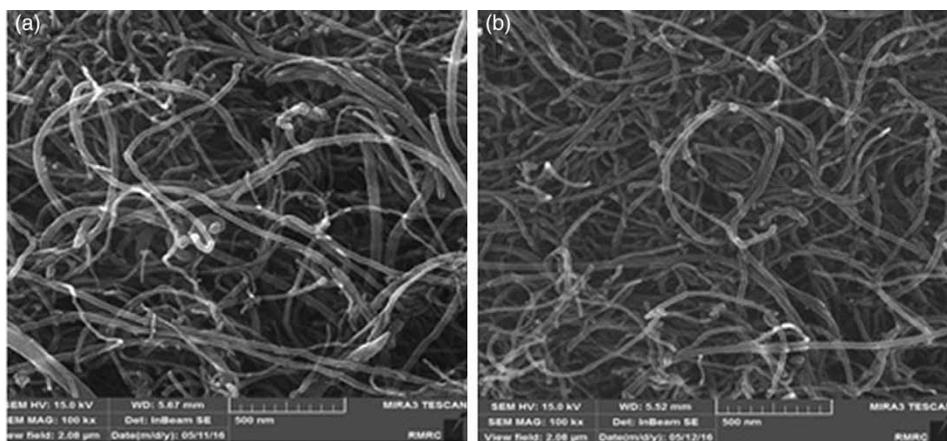


Figure 1 | SEM of MWCNTs (a) before and (b) after caffeine adsorption.

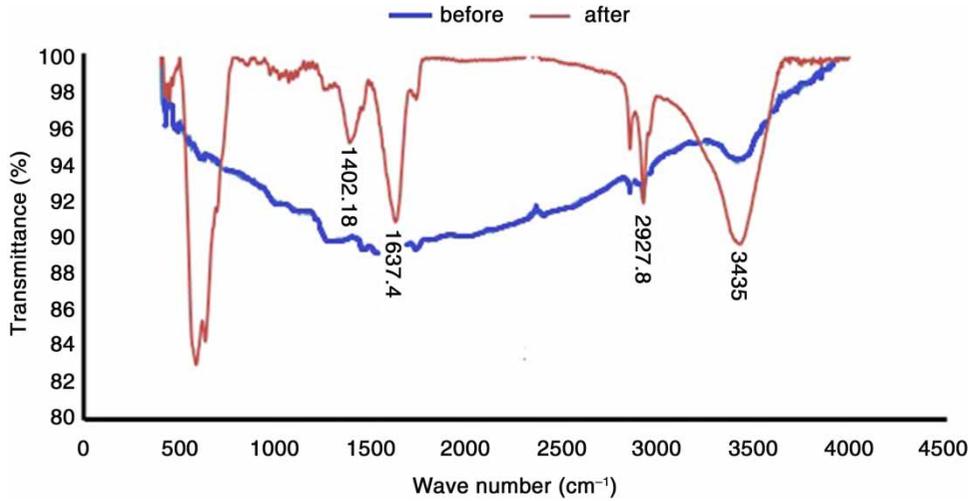


Figure 2 | FTIR spectra of MWCNTs before and after caffeine adsorption.

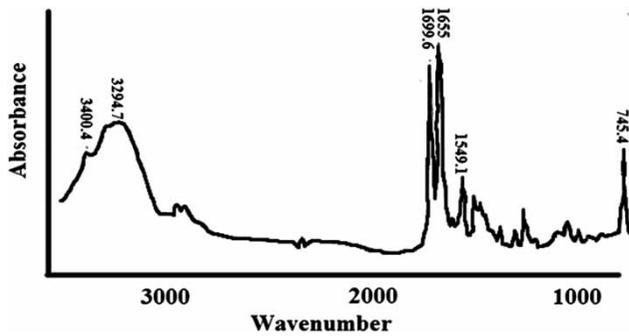


Figure 3 | FTIR spectra of standard caffeine.

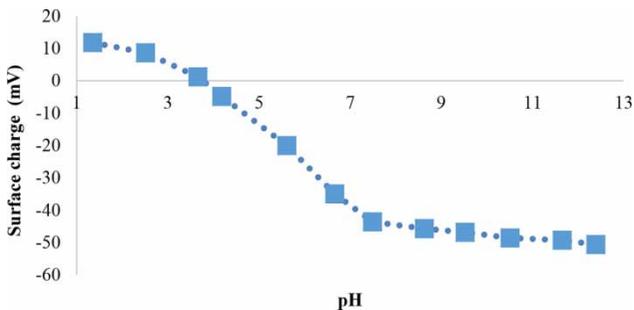


Figure 4 | Zeta potential of MWCNTs as a function of pH.

and that both interactions depend on the characteristics of the adsorbent and adsorbate, as well as the solution chemical properties (Moreno-Castilla 2004). As a similar substance, MWCNTs can be considered effective in removing organic contaminants.

Multiple polynomial regression study

Training the model

First, 86 samples of the data set were randomly selected for building the model then the remainder were entered and ran the model. The general equation of multiple polynomial regression with interaction effects is presented in Equation (1).

$$\begin{aligned}
 q = & b_1t + b_2C + b_3M + b_4pH + b_5T + b_6t^2 + b_7C^2 + b_8M^2 \\
 & + b_9pH^2 + b_{10}T^2 + b_{11}t^3 + b_{12}C^3 + b_{13}M^3 + b_{14}pH^3 \\
 & + b_{15}T^3 + b_{16}t^*C + b_{17}t^*M + b_{18}t^*pH + b_{19}t^*T \\
 & + b_{20}C^*M + b_{21}C^*pH + b_{22}C^*T + b_{23}M^*pH + b_{24}M^*T \\
 & + b_4pH^*T \quad (5)
 \end{aligned}$$

As can be seen in Equation (5), this model contains linear effects (t , C , M , pH , T), quadratic effects (t^2 , C^2 , M^2 , pH^2 , T^2), cubic effects (t^3 , C^3 , M^3 , pH^3 , T^3), and interaction effects (t^*C , t^*M , t^*pH , t^*T , C^*M , C^*pH , C^*T , M^*pH , M^*T , pH^*T). However, not all these effects are significant and step-by-step the non-effective parameters were removed. The final model has parsimonious parameters and maximum accuracy. Following on from this, we used Note 1 and 2 to determine the final model:

Note 1 – In polynomial regression when a cubic effect is significant, we must assume linear and quadratic effects in the model (significant or not significant). Also when a

quadratic effect is significant, we must assume linear effect in model (significant or not significant).

Note 2 – In regression with interaction effects, when an interaction effect is significant, we must assume both linear effects in model (significant or not significant).

The following terms could not be estimated and were removed: (t^*M , t^*pH , t^*T , C^*M , C^*pH , C^*T , M^*pH , M^*T , pH^*T , and T^3). Table 2 shows the results of the first step for multiple polynomial regression with interaction effects. As can be seen in Table 3, some of parameters were not significant (P -value >0.05). By considering Notes 1 and 2, the factors given in Table 3 were candidates for removal. Then the most not significant term (more P -value) in the first run (pH^3) was removed and the operation run again. The results of the second step for multiple polynomial regression

with interaction effects are given in Table 4. The factors given in Table 5 were the candidates for removal. The process was continued step-by-step to the point that there were only significant factors. Table 6 shows the results of the last run. As such, the cubic effect of t and C (t^3 , C^3), the quadratic effects of t , C and M (t^2 , C^2 and M^2), the linear effects of t , C and M and the interaction effect of t and C (t^*C) remained in model. Therefore, the final regression equation was as follows:

$$q = 1.191t + 0.3047C - 207.3M - 0.05659t^2 + 0.000054C^2 + 711.9M^2 - 0.006995t^*C + 0.001068t^3 - 0.000000C^3 \quad (6)$$

For the final model (Equation (6)), R^2 , R_{adj}^2 and R_{pred}^2 were obtained as 0.998, 0.9979 and 0.9973, respectively.

Table 2 | First step of regression analysis of variance

Source	P-value
t	0.000
C	0.000
M	0.000
pH	0.972
T	0.994
t^2	0.000
C^2	0.146
M^2	0.057
pH^2	0.982
T^2	0.989
t^*C	0.000
t^3	0.000
C^3	0.026
M^3	0.425
pH^3	0.991

Table 3 | Not significant parameters

Source	P-value
T^2	0.989
C^3	0.026
M^3	0.425
pH^3	0.991

Table 4 | Second step of regression analysis of variance

Source	P-value
t	0.000
C	0.000
M	0.000
pH	0.918
T	0.994
t^2	0.000
C^2	0.143
M^2	0.056
pH^2	0.914
T^2	0.989
t^*C	0.000
t^3	0.000
C^3	0.026
M^3	0.422

Table 5 | Not significant parameters

Source	P-value
pH^2	0.914
C^3	0.026
M^3	0.422
T^2	0.989

Table 6 | Significant parameters in last run

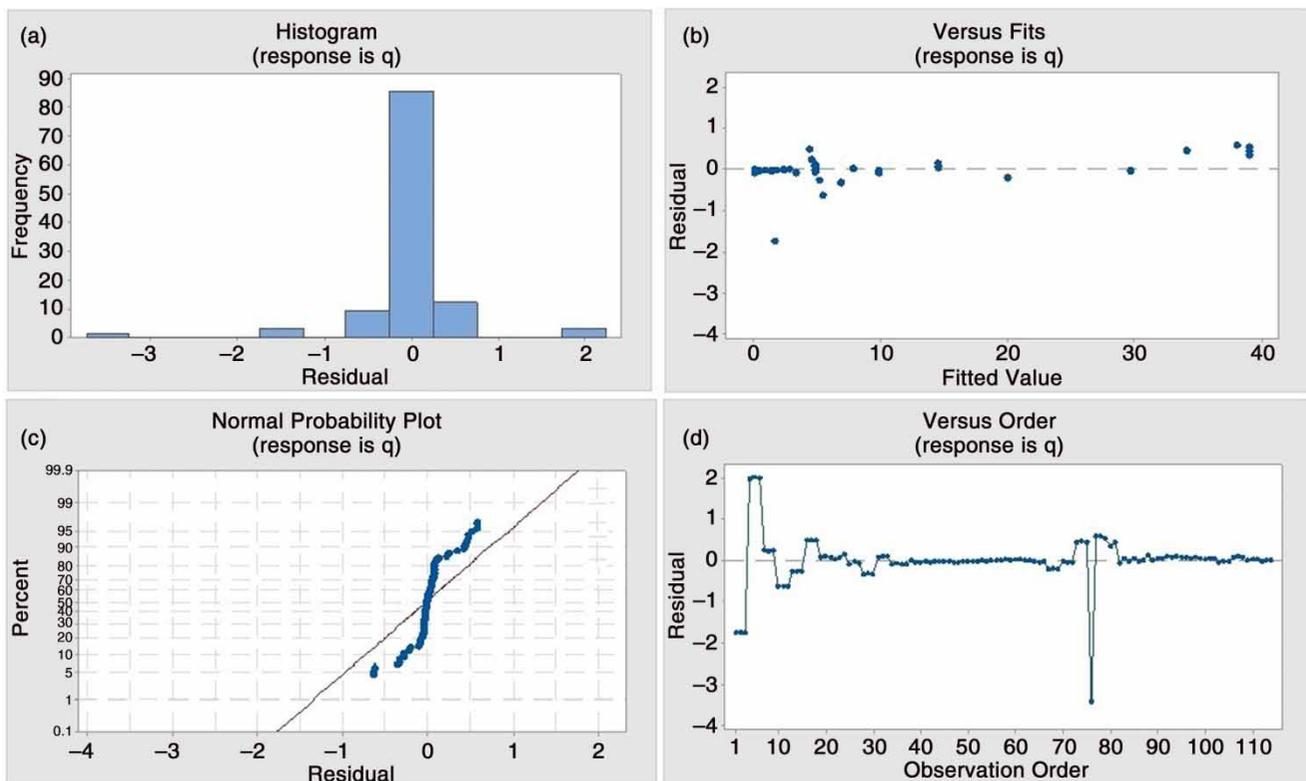
Source	P-value
t	0.000
C	0.000
M	0.000
t ²	0.000
C ²	0.093
M ²	0.000
t*C	0.000
t ³	0.000
C ³	0.013

Also, the RMSE of Equation (6) is 0.5987. Figure 5 shows the residual plots of the final model. This figure is applied to test the goodness of fit in multiple polynomial regression. Testing residual plots reveals that the usual least squares assumptions are being met. Figure 5(a) shows the histogram of the residuals, and whether the variance is normally distributed. An almost symmetric bell-shaped histogram of

Figure 5(a) displays the normality assumption is likely to be true. Figure 5(b) shows the residuals versus fits indicating that the residuals have a nearly constant variance. Figure 5(c) shows the normal probability plot of the residuals. As can be seen from Figure 5(c), the plot is nearly linear assuming that the error terms are normally distributed. Figure 5(d) shows the residuals versus order of data assuming that the residuals are uncorrelated with each other.

Testing the model

Twenty-eight independent data were selected for testing the Equation (6). Using the input data (t, C and M) in Equation (6), the adsorption capacity of caffeine, q (mg g^{-1}), was calculated. The linear regression was applied between measured adsorption capacity (q_m) and predicted adsorption capacity (q_p) of caffeine. Comparison of the q_p by multiple polynomial regression and q_m is presented in Figure 6. Also, the confidence limits at the 5% level, based on the distribution of points around the fitted line, indicate

**Figure 5** | (a) Histogram of the residuals; (b) residuals versus fits; (c) normal probability plot of residuals; and (d) residuals versus order of data.

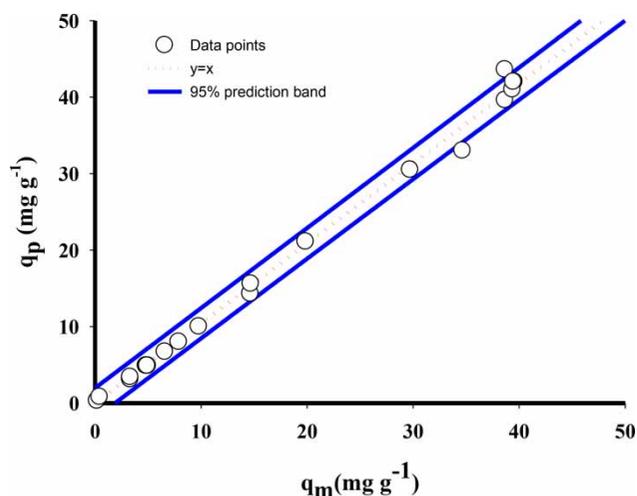


Figure 6 | Comparison of the q predicted by multiple polynomial regression and observed values.

an excellent reliability to predict adsorption capacity of caffeine. As Table 7 shows, the various tests results indicated that Equation (6) with inputs of t , C and M accomplished best in predicting q_p .

Results of the test data set show that Equation (6) slightly overestimated (MRE was positive). The R^2 and NRMSE of Equation (6) were calculated to be 0.9961 and 0.0916, respectively, showing the excellent performance of multiple polynomial regression in predicting q_p . As mentioned in the above section, the pH and T were removed in final model, therefore, the pH and temperature have no significant effect on adsorption capacity of caffeine in a batch mode experiment, and so the number of experiments could be reduced based on the results of the models. In this respect, for three independent variables (t , C and M), three structures were built. The statistical indices and performance of the three structures as compared with the best model (Equation (6)) are presented in Table 7.

For all linear regressions observed in Table 7, the distances from the origin are not significant at the 5% level

and are considered zero. Based on Table 7, q_p values of Equation (6), Structure1, Structure2 and Structure3 were higher than the measured adsorption capacity of caffeine, where the ratios of $q_{E(6)}/q_m$, $q_{Structure1}/q_m$, $q_{Structure2}/q_m$ and $q_{Structure3}/q_m$ were 1.05, 1.07, 1.09 and 1.25, respectively. Structure1 indicated good performance, with value of NRMSE equal to 0.1544. Performance of Structure2 was fair, with NRMSE equal to 0.2171, but when Structure3 was applied, results indicated that this equation gives poor prediction of the adsorption capacity of caffeine. To determine the sensitivity of the model to each independent variable, the variation of NRMSE was investigated as compared with the best model by eliminating each operating variable from the model. As shown in Table 7, greatest increase in NRMSE was due to C , followed by M , and then t . The most efficient variable is the initial concentration of caffeine, the NRMSE amount was markedly raised (0.0916 to 0.4731) when this parameter was removed from Equation (6). Similar outcomes have been demonstrated in prior studies (Jain et al. 2009; Mousavi et al. 2012). In fact, the greater removal efficiency of caffeine was observed at the lower concentration of caffeine due to the accessibility of more unoccupied active sites. However, the removal efficiency of caffeine decreased on raising initial caffeine concentration due to saturation of exchangeable sites of the MWCNTs. Therefore, the removal of caffeine by MWCNTs is more dependent on C .

The second most efficient parameter is mass of adsorbent, where the NRMSE amount raised from 0.0916 to 0.1544. Similarity, the third most efficient parameter is contact time, where the NRMSE amount raised from 0.0916 to 0.2171. This can be attributed to the short time of equilibrium and caused less effective caffeine adsorption by MWCNTs. In fact, equilibrium is reached after 5 min for caffeine (data not presented). The removal efficiency increases with time in the first 5 min and then the adsorption curve

Table 7 | Statistical performance evaluation criteria for the final models

Model	Input	MRE	R^2	NRMSE	Performance	Equation
Equation (6) (best model)	t , C and M	0.035	0.9961	0.0916	Excellent	$q_p = 1.05q_m$
Structure1	C and M	0.069	0.975	0.1544	Good	$q_p = 1.07q_m$
Structure2	C and t	0.18	0.929	0.2171	Fair	$q_p = 1.09q_m$
Structure3	M and t	0.814	0.4873	0.4731	Poor	$q_p = 1.25q_m$

reached equilibrium after this time. However, pH and temperature were not significant variables and were dropped from the final model which confirmed the applicability of MWCNTs for treatment of surface and ground water in the broad range of these parameters. This study demonstrated that the multiple polynomial regression could be accurate and a faster alternative to the available difficult and time-consuming models for q prediction.

CONCLUSION

In this research, the adsorption capacity of caffeine by MWCNTs is studied using multiple polynomial regression. The influences of C, M, t, T and pH on q (mg g^{-1}) were studied. The MWCNTs were specified by SEM, FTIR spectroscopy and point of zero charge. The results confirmed that the MWCNTs have a high capacity to uptake caffeine from the wastewater. Results show that multiple polynomial regression could give excellent fit to the observation adsorption capacity. The increment of input variables from only 'C + M + t' to 'C + M + t + pH + T' did not show a significant effect on the removal efficiency of caffeine from aqueous media and consequently did not affect the model. The results also indicated that C was more important in q prediction, relative to M and t. The suggested technique is easy to operate, accurate, rapid and needs less computational time.

REFERENCES

- Álvarez-Torrellas, S., Rodríguez, A., Ovejero, G., Gómez, J. M. & García, J. 2016 Removal of caffeine from pharmaceutical wastewater by adsorption: influence of NOM, textural and chemical properties of the adsorbent. *Environmental Technology* **37**, 1618–1630.
- Amiri, M. J., Abedi-Koupai, J., Eslamian, S. S., Mousavi, S. F. & Arshadi, M. 2013a Modelling Pb(II) adsorption based on synthetic and industrial wastewaters by ostrich bone char using artificial neural network and multivariate non-linear regression. *International Journal of Hydrology Science and Technology* **3**, 221–240.
- Amiri, M. J., Abedi-Koupai, J., Eslamian, S. S., Mousavi, S. F. & Hasheminejad, H. 2013b Modeling Pb(II) adsorption from aqueous solution by ostrich bone ash using adaptive neural-based fuzzy inference system. *Journal of Environmental Science and Health, Part A* **48**, 543–558.
- Arshadi, M., Amiri, M. J. & Mousavi, S. 2014 Kinetic, equilibrium and thermodynamic investigations of Ni(II), Cd(II), Cu(II) and Co(II) adsorption on barley straw ash. *Water Resources and Industry* **6**, 1–17.
- Ayranci, E., Hoda, N. & Bayram, E. 2005 Adsorption of benzoic acid onto high specific area activated carbon-cloth. *Journal of Colloid and Interface Science* **284**, 83–88.
- Balistrieri, L. S. & Murray, J. W. 1981 The surface chemistry of goethite ($\alpha\text{-FeOOH}$) in major ion seawater. *American Journal of Science* **281**, 788–806.
- Chen, Y. H., Huang, Y. H., Wen, C. C., Wang, Y. H., Chen, W. L., Chen, L. C. & Tsay, H. J. 2008 Movement disorder and neuromuscular change in Zebrafish embryos after exposure to caffeine. *Neurotoxicology and Teratology* **30**, 440–447.
- Gamze Turan, N., Mesci, B. & Ozgonenel, O. 2011a The use of artificial neural networks (ANN) for modeling of adsorption of Cu(II) from industrial leachate by pumice. *Chemical Engineering Journal* **171**, 1091–1097.
- Gamze Turan, N., Mesci, B. & Ozgonenel, O. 2011b Artificial neural network (ANN) approach for modeling Zn(II) adsorption from leachate using a new biosorbent. *Chemical Engineering Journal* **173**, 98–105.
- Ghaedi, M. & Kokhdan, S. N. 2012 Oxidized multiwalled carbon nanotubes for the removal of methyl red (MR): kinetics and equilibrium study. *Desalination and Water Treatment* **49**, 317–325.
- Ghaedi, M., Shokrollahi, A., Hossainian, H. & Kokhdan, S. N. 2011 Comparison of activated carbon and multiwall carbon nanotubes for efficient removal of eriochrome cyanine R (ECR): kinetic, isotherm, and thermodynamic study of the removal process. *Journal of Chemical & Engineering Data* **56**, 3227–3235.
- Glassmeyer, S. T., Furlong, E. T., Kolpin, D. W., Cahill, J. D., Zaugg, S. D., Werner, S. L., Meyer, M. T. & Keryak, D. D. 2005 Transport of chemical and microbial compounds from known wastewater discharges: potential for use as indicators of human fecal contamination. *Environmental Science & Technology* **39**, 5157–5169.
- Hosseini, S. J., Kokhdan, S. N., Ghaedi, A. M. & Moosavian, S. S. 2011 Comparison of multiwall carbon nanotubes and activated carbon for efficient removal of methyl orange: kinetic and thermodynamic investigation. *Fresenius Environmental Bulletin* **20**, 219–234.
- Iijima, S. 1991 Helical microtubules graphitic carbon. *Nature* **354**, 56–58.
- Jain, M., Garg, V. K. & Kadirvelu, K. 2009 Chromium (VI) removal from aqueous system using Heliantus annuus (sunflower) stem waste. *Journal of Hazardous Materials* **162**, 365–372.
- Jamieson, P. D., Porter, J. R. & Wilson, D. R. 1991 A test of computer simulation model ARC-WHEAT1 on wheat crops grown in New Zealand. *Field Crops Research* **27**, 337–350.

- Ma, H. Y., Burger, C., Hsiao, B. S. & Chu, B. 2012 Highly permeable polymer membranes containing directed channels for water purification. *ACS Macro Letters* **1**, 723–726.
- Moreno-Castilla, C. 2004 Adsorption of organic molecules from aqueous solutions on carbon materials. *Carbon* **42**, 83–94.
- Mousavi, S. F., Esteki, M., Mostafazadeh-Fard, B., Dehghani, S. & Khorvash, M. 2012 Linear and nonlinear modeling for predicting nickel removal from aqueous solutions by dried sunflower stalks. *Environmental Engineering Science* **29**, 765–775.
- Pavel, I., Szeghalmi, A., Moigno, D., Cinta, S. & Kiefer, W. 2003 Theoretical and pH dependent surface enhanced Raman spectroscopy study on caffeine. *Biopolymers* **72**, 25–37.
- Pollack, K., Balazs, K. & Ogunseitan, O. 2009 Proteomic assessment of caffeine effects on coral symbionts. *Environmental Science & Technology* **43**, 2085–2091.
- Razi, M. A. & Athappilly, K. 2005 A comparative predictive analysis of neural networks (NNs), nonlinear regression and classification and regression tree (CART) models. *Expert Systems with Applications* **29**, 65–74.
- Tan, C. W., Tan, K. H., Ong, Y. T., Mohamed, A. R., Zein, S. H. S. & Tan, S. H. 2012 Energy and environmental applications of carbon nanotubes. *Environmental Chemistry Letters* **10**, 265–273.
- Wang, Y., Ma, J., Zhu, J., Ye, N., Zhang, X. & Huang, H. 2016 Multi-walled carbon nanotubes with selected properties for dynamic filtration of pharmaceuticals and personal care products. *Water Research* **92**, 104–112.
- Willmott, C. J., Rowe, C. M. & Mintz, Y. 1985 Climatology of terrestrial seasonal water cycle. *Journal of Climate* **5**, 589–606.
- Yetilmezsoy, K. & Demirel, S. 2008 Artificial neural network (ANN) approach for modeling of Pb(II) adsorption from aqueous solution by Antep pistachio (*Pistacia Vera* L.) shells. *Journal of Hazardous Materials* **153**, 1288–1300.
- Yu, J. G., Zhao, X. H., Yang, H., Chen, X. H., Yang, Q., Yu, L. Y., Jiang, J. H. & Chen, X. Q. 2014 Aqueous adsorption and removal of organic contaminants by carbon nanotubes. *Science of the Total Environment* **482–483**, 241–251.

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