

Organic matter capture by a high-rate inoculum-chemostat and MBBR system

Hadi Abbasi, Charles Élysée, Marc-André Labelle, Edith Laflamme, Alain Gadbois, Antoine Laporte, Peter L. Dold and Yves Comeau

ABSTRACT

The main objective of this study was to develop an innovative process to maximize the bio-transformation of colloidal and soluble biodegradable matter (CS_B) into particulate matter (X_B) for energy recovery via methane production. Two configurations were studied: (1) high-rate moving bed bioreactor (HR-MBBR) and (2) inoculum-chemostat (IC) system consisting of a very HR-MBBR inoculating a continuous flow stirred-tank reactor. The effect of hydraulic retention time (HRT), specific organic loading rate (SOLR), and dissolved oxygen (DO) level were determined using real wastewater at pilot scale. Results showed that in the HR-MBBR process, a very high CS_B bio-transformation efficiency (90%) was obtained in a wide range of SOLRs (2.0 to 5.5 g CS_B m⁻² d⁻¹) corresponding to an optimum HRT of 36 minutes. The IC process reached a maximum CS_B bio-transformation efficiency of 77%, at SOLRs ranging from 22 to 30 g CS_B m⁻² d⁻¹ at an HRT of 3.7 hours. The DO concentration in the HR-MBBR influenced the CS_B bio-transformation ratio, while the HRT and the SOLR were the dominant factors influencing this ratio in the IC process. Based on these results, the IC process could be an interesting alternative to high-rate systems towards obtaining energy positive/efficient from water resource recovery facilities.

Key words | chemostat, COD oxidation, high-rate MBBR, organic matter capture

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INTRODUCTION

Environmental protection requirements and energy demand are major factors driving the energy efficiency of water resource recovery facilities (WRRFs). Conventional processes, like activated sludge (AS), are widely used for wastewater treatment, but they require a significant amount of energy (Jimenez *et al.* 2015). Therefore, process optimization and innovative treatment strategies are

required to improve the energy balance and obtain cost-effective WRRFs (Metcalf & Eddy-Aecom 2014; Meerburg *et al.* 2015).

A central approach to obtain energy-positive WRRFs is to maximize the capture of organic matter for energy production via methanogenesis. The biodegradable organic matter consists of readily (RBCOD) and slowly biodegradable

(SBCOD) fractions (Henze *et al.* 2000; Melcer 2003). Readily biodegradable matter is composed of soluble (S_B) and colloidal (C_B) matter that can be oxidized or stored directly by heterotrophic bacteria under aerobic conditions and used for the growth of new heterotrophic biomass X_{OHO} via bio-transformation processes. Slowly biodegradable matter (mostly particulate matter, X_B and ordinary heterotrophic organisms, X_{OHO}) requires conversion into a readily biodegradable form by hydrolysis prior to absorption and utilization. Thus, optimizing aeration, minimizing hydrolysis, minimizing oxidation of particulate matter, and capturing biodegradable organic matter to be sent to anaerobic digestion can improve the energy efficiency of WRRFs (Ødegaard 2000; Jimenez *et al.* 2015).

High-rate biological treatment profits from the high bacterial activity under high food-to-microorganism ratios and low solid retention times (SRTs) with relatively short hydraulic retention times (HRTs) resulting in the maximization of bio-transformation and capture of organic matter from wastewater (Grady *et al.* 2011; Jimenez *et al.* 2015). The high-rate moving bed bioreactor (HR-MBBR) is a promising process which is successfully used for organic matter recovery at low HRTs (30–90 min) while still maintaining a high COD removal efficiency (80–85%) (Ødegaard 2000). Biomass is grown in such HR processes, transforms C_B and S_B (CS_B) into X_{OHO} , minimizing the oxidation of X_B while increasing the production of X_{OHO} , thus maximizing the energy generation potential (Jimenez *et al.* 2015; Brosseau 2015).

The main objective of this study was to develop an innovative process combining an HR-MBBR and an AS process to maximize the bio-transformation of CS_B into X_B for energy recovery via methane production. For this purpose, two pilot-scale treatment configurations, including a high-rate MBBR (HR-MBBR) in parallel with a very HR-MBBR acting as an inoculum and an AS chemostat (IC) system, were tested to address the following specific objectives:

1. determine the operational parameters (HRT, specific organic loading rate (SOLR), dissolved oxygen (DO)) to maximize the performance of each treatment process;
2. maximize the bio-transformation of CS_B into X_B to allow the capture of X_B to maximize methane production.

MATERIALS AND METHODS

Pilot plant setup and configurations

The pilot plant, comprising (a) an HR-MBBR (1.6 m³) and (b) a very HR-MBBR inoculum (0.4 m³) followed by a chemostat (4.7 m³) (Figure 1), was installed at the Repentigny municipal WRRF, Quebec. The raw wastewater influent containing about 10% industrial loading was subjected to 6 mm screening, fat and grease removal and grit removal prior to being fed to the pilot plant trains. The wastewater characteristics and operating conditions (OC) of the HR-MBBR and IC are presented in Table 1.

Additional screening was provided by another 6 mm punched hole strainer which was connected at the inlet of both systems to remove trash and which was cleaned manually every 2 days. The HR-MBBR and inoculum were filled with the carrier type K5 from AnoxKaldnesTM with a specific surface area of 800 m²/m³. All reactors were completely mixed and were equipped with fine and coarse bubble aeration systems, and a mechanical mixer. Probes (Hach) connected to the automation system (PLC), were used for real-time monitoring of DO, pH, and temperature.

Aeration

The aeration system in the HR-MBBR process provided 16.3 ± 1.2 m³/h provided through coarse (1/3) and fine (2/3) bubbles to ensure proper aeration and media mixing. The aeration system in the IC system provided 1.5 ± 0.2 and 2.5 ± 1.3 m³/h in the inoculum and chemostat processes, respectively, via fine bubble diffusers.

Sampling and analytical methods

The influent to each process was sampled two to five times per week. Multiple grab samples (taking into account the HRT) from the influent were mixed together to obtain a homogeneous composite sub-sample. Total and soluble COD, total and volatile suspended solids were analyzed at each sampling point according to Standard Methods 5220D (APHA 2012). Filtered COD was determined using both 1.2 µm glass microfiber filters (Whatman[®] 934-AHTM, GE Healthcare Life

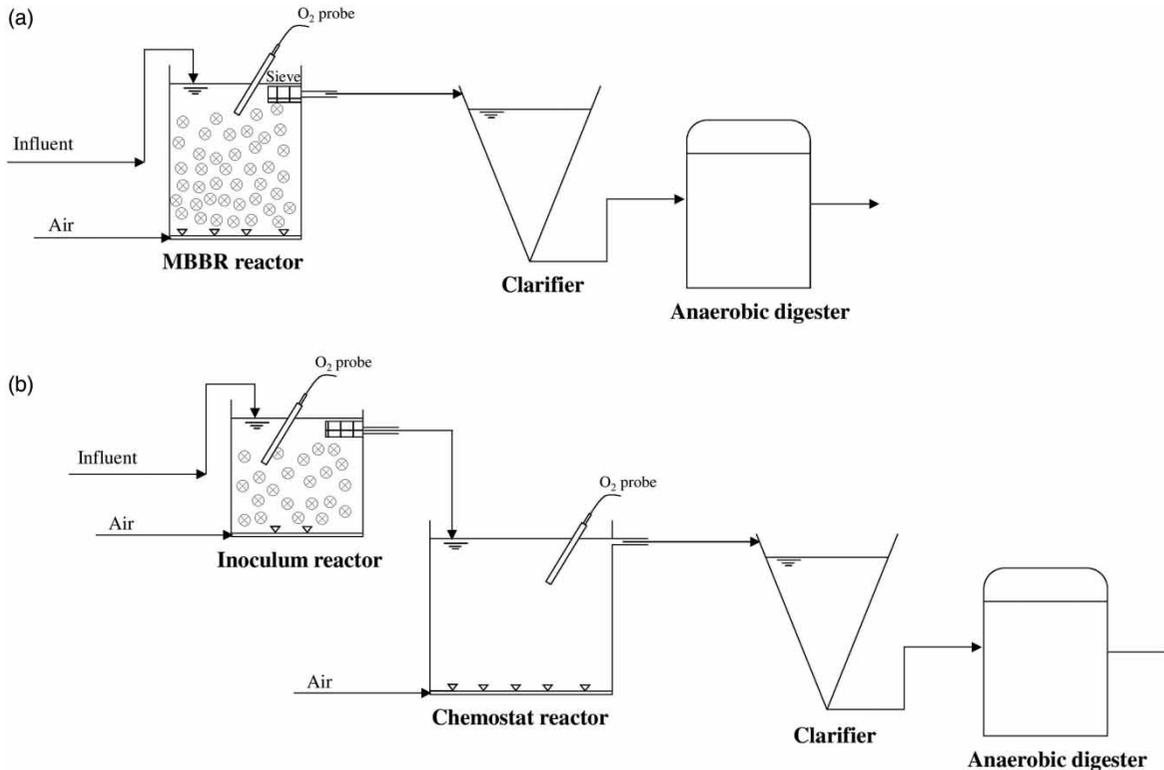


Figure 1 | Configuration of the (a) HR-MBBR and (b) IC treatment systems.

Sciences, GBR) and 0.45 μm cellulose membranes (MF-Millipore™, EMD Millipore). Flocculated-filtered COD (ffCOD) was measured using the method developed by Mamais & Jenkins (1993). The COD fractions characterized were thus particulate COD ($X_{\text{COD}} > 1.2 \mu\text{m}$), colloidal and soluble COD ($\text{CS}_{\text{COD}} < 1.2 \mu\text{m}$), and soluble COD ($\text{ffCOD} = \text{S}_{\text{COD}} < 0.45 \mu\text{m}$). Colloidal COD (C_{COD}) fraction was calculated from the difference between CS_{COD} and S_{COD} . The colloidal and soluble unbiodegradable fraction (S_U) was considered to be the typical 5% of the total COD (EnviroSim Associates Ltd 2016). The following formula was used to calculate the C_U , C_B , and S_B , according to S , C , CS , and S_U (given above) values:

$$C_U = S_U \times (\text{CS}/\text{S} - 1) \quad (1)$$

$$C_B = C_{\text{COD}} - C_U \quad (2)$$

$$S_B = \text{S}_{\text{COD}} - S_U \quad (3)$$

The DO was measured with a portable DO-meter (HQ40d, Hach Company) and an LDO® probe (Hach Company).

The biofilm mass was measured every week by collecting carriers (20 carriers per sampling event) dried at 105 °C overnight and weighed. The carriers were then soaked in 6% NaOH for 30 min to recover the biofilm from the carrier surface, after which the carriers were scraped clean and dried again at 105 °C overnight. The difference between the dry weight of the carriers before and after cleaning represented the mass of biofilm on the carriers. The amount of biofilm per square meter of protected surface area of carriers ($\text{g TSS}/\text{m}^2$) was determined by dividing the obtained total solids (TS) of the detached biofilm over the protected surface area of the number of carriers sampled (Andreottola et al. 2000, 2003). Considering a protected surface area of 23 $\text{cm}^2/\text{carrier}$ allowed determination of the specific biofilm concentration in g/m^2 .

Table 1 | Influent and process and operating characteristics for the pilot-scale reactors at different OC

Parameter	Units	HR-MBBR					Inoculum					Chemostat				
		OC1	OC2	OC3	OC4	OC5	OC1	OC2	OC3	OC4	OC5	OC1	OC2	OC3	OC4	OC5
Influent																
Q	m ³ /h	1.8	2.6	3.8	1.8	3.8	2.0	1.35	1.0	2.0	2.0	2.0	1.4	1.0	2.0	2.0
Total COD	mg/L	403 ± 43	432 ± 77	409 ± 59	400 ± 59	390 ± 100	440 ± 81	454 ± 22	455 ± 39	381 ± 36	370 ± 24	429 ± 48	440 ± 24	429 ± 41	380 ± 23	271 ± 10
Colloidal COD	mg/L	26 ± 13	26 ± 14	37 ± 8	31 ± 14	21 ± 15	27 ± 13	16 ± 9	52 ± 13	24 ± 14	29 ± 12	21 ± 13	10 ± 4	32 ± 13	8 ± 5	29 ± 5
Soluble COD	mg/L	61 ± 8	68 ± 15	69 ± 10	60 ± 9	61 ± 14	64 ± 14	68 ± 8	73 ± 7	61 ± 4	56 ± 11	55 ± 12	53 ± 6	50 ± 3	53 ± 7	42 ± 2
Process and operating characteristics																
Liquid volume	m ³	1.6	1.6	1.6	1.6	1.7	0.42	0.36	0.36	0.43	0.43	4.7	4.7	4.7	4.7	4.7
HRT	min	54	36	25	54	25	13	16	22	13	13	141	209	282	141	141
COD loading	kg COD/d	17 ± 2	29 ± 5	39 ± 6	14 ± 4	38 ± 9	21 ± 4	14 ± 1	11 ± 1	17 ± 2	14 ± 1	21 ± 2	14 ± 1	10 ± 1	239 ± 23	182 ± 6
Fill volume fraction	m ³ /m ³	50	50	35	50	50	13	16	16	25	45	-	-	-	-	-
SOLR ^a	g m ⁻² d ⁻¹	26 ± 3	46 ± 9	85 ± 12	23 ± 6	59 ± 14	474 ± 87	319 ± 16	237 ± 20	238 ± 1	196 ± 1	-	-	-	-	-
Temp	°C	17-22	17-22	18-21	17-18	18-21	17-22	17-22	18-21	19-20	16-17	17-22	17-22	18-21	19-20	16-17
DO	mg/L	2-4	3-4	3-4	1.5-2	3-4	5-6	5-6	5-6	4-6	4-6	6-7	6-7	6-7	6-7	6-7

^aThe specific organic loading rate (SOLR) was calculated based on total COD.

Statistical analysis

Statistical comparisons between the HR-MBBR and IC treatment efficiencies were conducted using the t-test function in Microsoft Excel 2013 with the least significant difference of $P < 0.05$.

RESULTS AND DISCUSSION

The effect of HRT, SOLR, media fill volume fraction, and oxygen uptake rate (OUR) was considered in the following sections for maximizing the production of biodegradable sludge, based on the maximization of the removal efficiency of CS_B (bio-transformation of CS_B into X_{COD}) as well as the minimization of biodegradable particulate matter (X_B) hydrolysis.

A summary of the pilot-scale HR-MBBR and IC effluent characteristics of the five OC is presented in Table 2.

No significant nitrification occurred, as expected under such high-rate conditions, as shown by the very low concentration of nitrate (0.1 mg N/L) in the effluent of the HR-MBBR and IC processes.

Effect of HRT, SOLR, and DO on bio-transformation of CS_B and hydrolysis X_{COD}

The effect of HRT on the bio-transformation of the C_B and S_B , and the hydrolysis of X_{COD} in the HR-MBBR at OC OC1, OC2, and OC5 based on the HRT are shown in Figure 2(a) and Table 2. C_B and S_B bio-transformation increased from $75 \pm 5\%$ to $83 \pm 6\%$ by increasing HRT from 25 min to 54 min. The bio-transformation of C_B and S_B into X_B showed no significant difference at HRTs longer than 36 min and reached a plateau at $85 \pm 6\%$ (below an SOLR 36 ± 6 kg COD m⁻² d⁻¹).

The same tendency also has been observed between HRT and CS_B bio-transformation by Brosseau et al. (2016) and Aygun et al. (2008); however, the bio-transformation efficiencies were systematically different due to influent COD concentration, available surface area for biofilm growth, and HRT in their experiments.

A lower value of X_{COD} and COD removal efficiency was observed at higher HRT (36 min and 54 min), probably due

Table 2 | Summary of OC, effluent characteristics, and process performance for the pilot-scale HR-MBBR and IC

Parameter	Symbol	Units	HR-MBBR					IC				
			OC1	OC2	OC3	OC4	OC5	OC1	OC2	OC3	OC4	OC5
OC												
COD loading	-	kg COD/d	17 ± 2	29 ± 5	39 ± 6	14 ± 4	38 ± 9	21 ± 2	15 ± 1	11 ± 1	18 ± 2	18 ± 7
HRT	-	min	54	36	25	54	25	154	225	304	154	154
Solids retention time	SRT	d	1.4	1.4	1.6	-	1.6	0.6	0.8	0.7	-	0.6
Biofilm												
Total suspended solids	TSS	mg/L	-	641 ± 5	868 ± 122	500 ± 64	1,168 ± 260	351 ± 68	357 ± 62	340 ± 60	119 ± 13	-
VSS/TSS ratio	f _{VT}	g VSS/g TSS	-	0.63 ± 0.02	0.77 ± 0.02	0.64 ± 0.03	0.63 ± 0.02	0.74 ± 0.03	0.70 ± 0.00	0.70 ± 0.02	0.69 ± 0.02	-
Effluent												
Total COD	COD	mg COD/L	313 ± 39	357 ± 58	335 ± 71	338 ± 92	307 ± 89	414 ± 55	371 ± 12	406 ± 95	389 ± 31	341 ± 56
Colloidal COD	C _{COD}	mg COD/L	12 ± 3	8 ± 4	17 ± 5	22 ± 6	6 ± 3	14 ± 8	8 ± 4	19 ± 11	10 ± 6	21 ± 7
Soluble COD	S _{COD}	mg COD/L	24 ± 4	29 ± 8	38 ± 8	33 ± 5	29 ± 9	40 ± 10	40 ± 10	38 ± 3	41 ± 5	40 ± 4
TSS	X _{TSS}	mg TSS/L	295 ± 67	327 ± 67	245 ± 80	316 ± 114	320 ± 141	364 ± 51	303 ± 63	256 ± 66	302 ± 54	319 ± 54
VSS	X _{VSS}	mg VSS/L	166 ± 25	206 ± 35	182 ± 47	184 ± 48	170 ± 52	218 ± 31	171 ± 25	185 ± 29	185 ± 27	170 ± 31
VSS/TSS ratio	f _{VT}	g VSS/g TSS	0.57 ± 0.07	0.63 ± 0.04	0.76 ± 0.07	0.60 ± 0.09	0.57 ± 0.13	0.60 ± 0.06	0.57 ± 0.07	0.74 ± 0.10	0.62 ± 0.04	0.59 ± 0.10
X _{COD} /VSS ratio	f _{CV}	g X _{COD} /g VSS	1.7 ± 0.1	1.6 ± 0.1	1.9 ± 0.1	1.5 ± 0.1	1.8 ± 0.1	1.7 ± 0.2	1.7 ± 0.2	1.9 ± 0.4	1.8 ± 0.1	1.7 ± 0.2
Alkalinity	S _{Alk}	mg CaCO ₃ /L	-	-	156 ± 10	-	162 ± 11	184 ± 15	-	193 ± 18	-	-
pH	-	-	7.3 ± 0.2	7.5 ± 0.2	7.4 ± 0.2	6.7 ± 0.2	7.3 ± 0.2	7.9 ± 1.0	7.5 ± 0.1	7.7 ± 0.4	8.1 ± 0.4	6.7 ± 0.1
Process performance												
CS _B biotransform efficiency	R _{CSB}	%	85 ± 6	86 ± 9	67 ± 4	64 ± 13	78 ± 13	56 ± 5	62 ± 5	74 ± 6	53 ± 6	40 ± 11
CS _B specific removal rate	SR _{CSB}	g CS _{COD} m ⁻² d ⁻¹	3 ± 1	6 ± 1	10 ± 2	3 ± 1	7 ± 2	39 ± 10	35 ± 4	28 ± 4	19 ± 7	13 ± 5
X _{COD,eff.} / (CS _B + X _{OHO}) _{inf.}	-	g X _{COD} /g BCOD	0.83 ± 0.15	0.86 ± 0.06	-	-	0.88 ± 0.12	1.01 ± 0.18	0.91 ± 0.12	0.95 ± 0.33	-	-

to the partial release of particulate matter from biofilms caused by abrasion at a long HRT (Hoang 2013). Minimal hydrolysis of particulate organic matter can be achieved at a low HRT, as can be achieved in a HR-MBBR process (Schubert *et al.* 2013).

Similarly, the effect of HRT on removal and bio-transformation of COD, X_{COD} , C_{B} , and S_{B} were evaluated through the IC process at OC OC1, OC2, and OC3 in Figure 2(b) and Table 2.

A positive correlation was observed between SC_{B} bio-transformation and HRT in the IC system due to the prolonged contribution of inoculum by transferring and establishing active biomass in the chemostat at higher HRT.

The CS_{B} bio-transformation efficiencies were $56\% \pm 5\%$, $62\% \pm 5\%$, and $74 \pm 6\%$ at HRTs of 154 min, 225 min, and 304 min, respectively, across the IC process. The concentration of COD and X_{COD} did not effectively change in the IC process at HRTs 154 and 304 min based on removal efficiency compared to HR-MBBR, due to the minimum effect of hydrolysis on particulate matter.

This phenomenon supported the results of Confer & Logan (1998), which found that hydrolysis rate is much more on the biofilm surface than at the surface of sloughed biofilm.

The overall efficiency of biotransformation of influent biodegradable organic matter into particulate matter across each process was also characterized by the ratio of effluent particulate COD to influent total biodegradable

COD. Results are presented in Table 2 as $X_{\text{COD,eff.}}/(CS_{\text{B}} + X_{\text{OHO}})_{\text{inf}}$. This fraction was lower across the IC process (0.96 ± 0.22 g X_{COD}/g BCOD) than the HR-MBBR process (0.86 ± 0.11 g X_{COD}/g BCOD), suggesting that less hydrolysis of particulate organic matter took place in the first one.

The effect of the SOLR on the removal of CS_{B} was also assessed in HR-MBBR and IC processes (Figure 3). The higher specific removal rates were attained as the SOLR was increased in both HR-MBBR and IC processes, whereas HRT and SOLR were identified as an important constraint on the bio-transformation (especially for IC process).

A maximum CS_{B} bio-transformation rate ($90 \pm 3\%$) in the HR-MBBR process was achieved at SOLR from 2.0 to 5.5 g $CS_{\text{B}} \text{ m}^{-2} \text{ d}^{-1}$, corresponding to an optimum HRT of 36 min. These values for the IC process reached in maximum the specific removal of $80 \pm 3\%$ corresponding to SOLR of 22 to 40 g $CS_{\text{B}} \text{ COD m}^{-2} \text{ d}^{-1}$ at an optimum HRT of 225 min.

The observed linear pattern between SOLR and CS_{B} removal efficiency was observed with the studies of Ødegaard (2000), Brosseau (2015), and Helness *et al.* (2005) in lab- and pilot-scale experiments with HR-MBBRs. Aygun *et al.* (2008) also demonstrated that by increasing the SOLR from 6 to 96 g $\text{COD m}^{-2} \text{ d}^{-1}$, the organic removal efficiency decreased from 95% to 45%.

In this context, Orantes & Gonzalez-Martinez (2003) established an asymptotic relationship between the mass of attached biofilm and SOLR, in which no further biomass is attached at high SOLR. Hence, at high SOLR, less biofilm

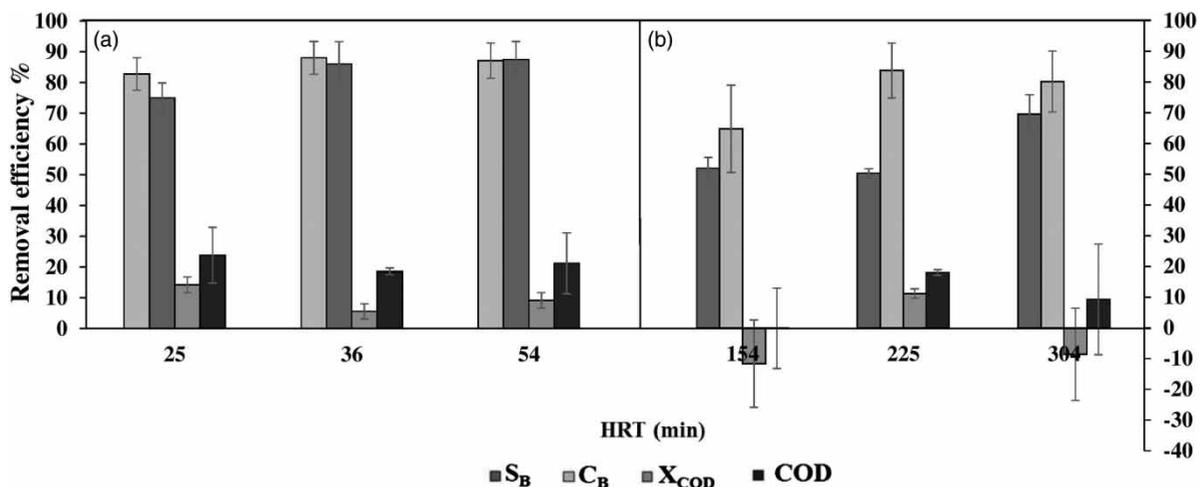


Figure 2 | Effect of HRT on the removal efficiency of COD fractions for the (a) HR-MBBR and (b) IC processes.

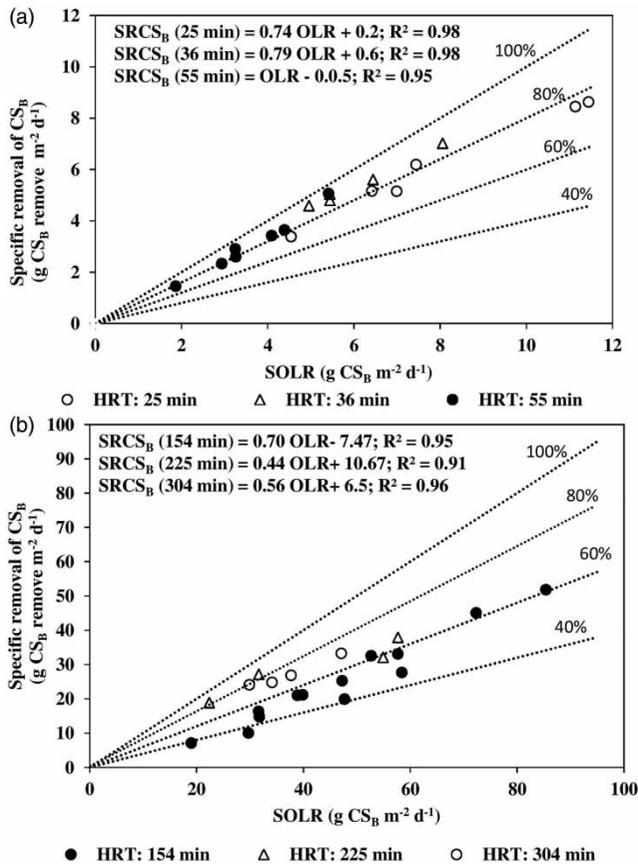


Figure 3 | Bio-transformation of CS_B into X_{COD} as a function of CS_B -SOLR at different HRTs: (a) HR-MBBR and (b) IC process.

can be established through the inoculum process and limited by short HRT, less contribution of inoculum could reasonably be expected to transfer active biomass into the chemostat.

Variation of DO during aerated and non-aerated periods in the HR-MBBR and chemostat was monitored, based on the OC of OC2 and OC1, respectively. The oxygen concentration dropped more rapidly in the HR-MBBR than in the chemostat when the aeration system was switched off for 3 minutes. Calculation of the OUR in two reactors indicated over five-fold higher OUR in the MBBR (50 ± 2 mg $O_2 L^{-1} h^{-1}$) than in the chemostat (10 ± 1.5 mg $O_2 L^{-1} h^{-1}$). In this context, the SOUR value across HR-MBBR and chemostat process was 55 ± 1 mg $O_2 g^{-1} VSS h^{-1}$ and 53 ± 6 mg $O_2 g^{-1} VSS h^{-1}$, respectively.

The highest OUR in the HR-MBBR can be correlated to the oxidation of more readily biodegradable matter

produced by hydrolysis of biofilm surface (Confer & Logan 1998) and the slowly biodegradable matter that results from lysis of decayed biomass in the HR-MBBR system, whereas the source of active biomass in the chemostat process was provided from inoculum continually with lower SRT (SRT_{IC-OC1} : 0.6 d and $SRT_{HR-MBBR-OC2}$: 1.4 d) and no further accumulation.

The removal efficiency of filterable biodegradable organic matter (CS_B) by the IC and HR-MBBR processes was about 75% and 85%, respectively, which corresponded to HRT of 141 min and 36 min.

Further tests were conducted to assess the effect of oxygen concentration on the bio-transformation rate in the HR-MBBR process. For this purpose, the DO concentration was changed from 1–2 mg O_2/L to 2–4 mg O_2/L during OC4 and OC1, respectively, for a duration of 1 week each.

The role of the DO concentration as an effective and sensitive parameter controlling the removal of S_B and C_B fractions, but not that of particulate COD, is illustrated in Figure 4 and Table 2. The maximum S_B and C_B removal efficiency, $86 \pm 7\%$ and $77 \pm 17\%$, respectively, was obtained during OC1. The S_B and C_B removal efficiency was significantly decreased to $67 \pm 20\%$ and $53 \pm 12\%$, respectively, as DO concentration was less than 2 mg O_2/L , indicating that the DO concentration (below 2 mg O_2/L) was a limiting factor in the HR-MBBR system. For an optimal COD removal, DO should be maintained higher than 2 mg O_2/L , as indicated by a 13% decline in COD removal when the

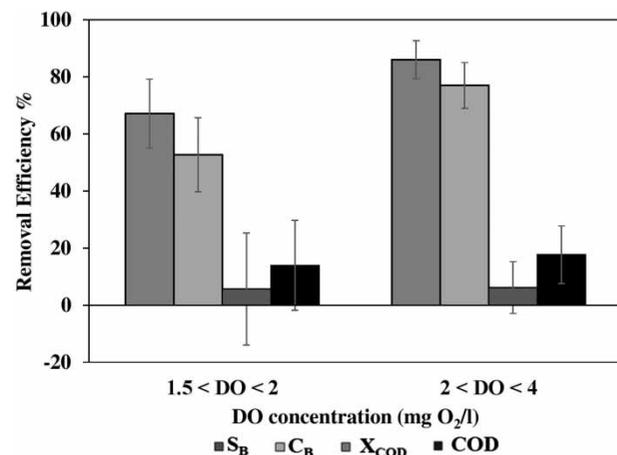


Figure 4 | Effect of DO on the COD removal efficiency of the different COD fractions across the HR-MBBR process (OC4 and OC1).

DO level was decreased from 2 to 1 mg/L while only a 6% increase in COD removal was observed with an increase in DO level from 2 to 6 mg/L (Wang et al. 2005).

Effect of media fill volume fraction on bio-transformation of CS_B

The effect of the two media fill volume fraction (OC3: 35% v/v, OC5: 50% v/v) on HR-MBBR treatment efficiency was assessed at HRT of 36 min. As the media fill volume fraction increased from 35% to 50% in the HR-MBBR, both S_B and C_B bio-transformation efficiency was increased from 76 ± 4 and 84 ± 6 to 89 ± 10 , 90 ± 8 , respectively (Figure 5(a) and Table 2). Similarly, Azizi et al. (2013) reported an effective treatment can be obtained by increasing media fill volume fraction up to 40% v/v, due to higher available surface area for biofilm growth. The removal of particulate and total COD decreased slightly by increasing the media fill volume fraction. Collision and attrition in the HR-MBBR reactor could lead to biofilm detachment from the outer surface and increase the total and particulate COD in the effluent due to the high volume of media and shear forces (Ødegaard 2000).

The effect of media fill volume fraction in the inoculum based on OC1, OC4, and OC5, aimed at transferring active biomass to the chemostat, demonstrated the opposite effect on bio-transformation of S_B and C_B in the IC process. Increasing the media fill volume fraction from 15% v/v to 45% v/v in the inoculum decreased the bio-transformation

of S_B and C_B from $58 \pm 4\%$ to $50 \pm 6\%$ and $69 \pm 14\%$ to $43 \pm 10\%$, respectively, in the IC process (Figure 5(b)). The media fill volume fraction ranging from 15% to 22% in the inoculum did not significantly affect the removal of particulate and total COD, while increasing the media fill volume fraction up to 45% significantly influenced the removal of total and particulate COD.

Higher media fill volume fraction (up to 45% v/v) may increase the development of active biomass in the inoculum (less biomass to be sloughed off the media) and may lead to less transferring of active biomass from inoculum to chemostat due to HRT constraints, therefore, less removal of S_B and C_B occurred in the chemostat.

Effect of OC on attached biofilm concentration

The attached biofilm growth concentration in the HR-MBBR was directly correlated to the SOLR (Figure 6).

Biofilm growth concentration in the HR-MBBR reactor reached a plateau of 18.0 ± 1.6 g TSS/m² at SOLR more than 8.0 ± 2.7 g CS_B m⁻² d⁻¹. In this context, a two-parameter exponential equation (R^2 : 0.95) showed the best fit to the biofilm concentration data. The concentration of attached biofilm during OC1 to OC5 was increased from 4.5 ± 2.6 g TSS/m² to 18.5 ± 1.2 g TSS/m² by increasing the SOLR from 3.1 ± 0.9 g CS_B m⁻² d⁻¹ to 15.6 ± 2.8 g CS_B m⁻² d⁻¹, respectively.

The attached biofilm concentration in inoculum after an increase from 6.0 ± 0.5 to 13.1 ± 0.8 g TSS/m² reached a

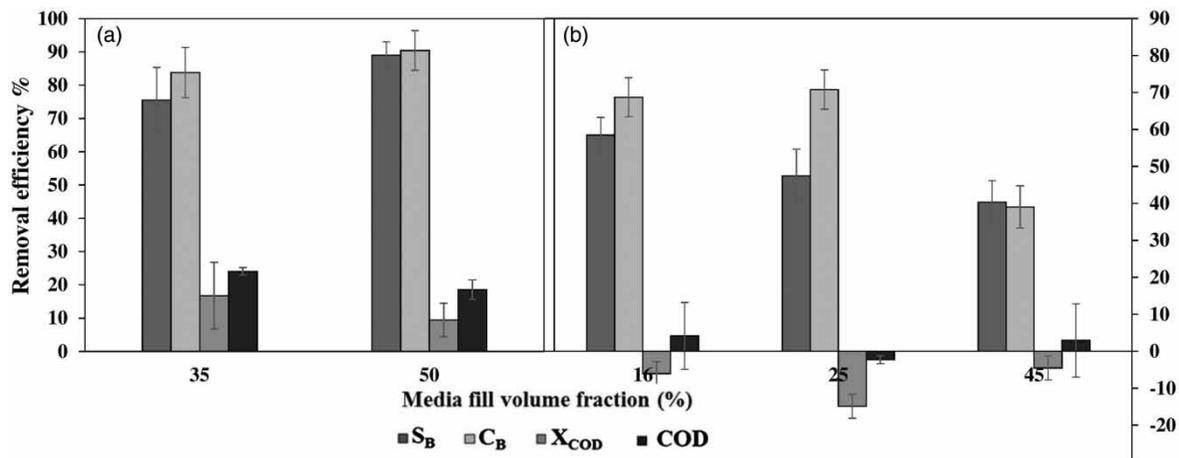


Figure 5 | Effect of the media fill volume fraction on the COD removal efficiency of COD fractions in the (a) HR-MBBR (OC3 and OC5) and (b) IC (OC1, OC4, and OC5) processes.

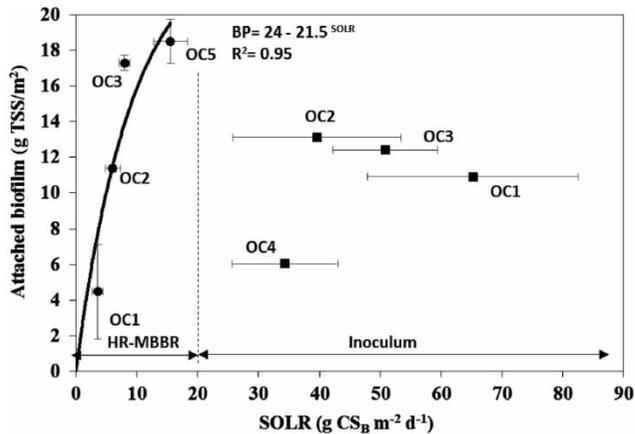


Figure 6 | Effect of SOLR on the attached biofilm concentration for different OC in the (a) HR-MBBR and (b) inoculum processes.

plateau with an average concentration of 11.7 ± 1.1 g TSS/m², while the SOLR ranged over 39.7 ± 13.8 g CS_B m⁻² d⁻¹. Moreover, under high SOLR in the inoculum, sloughing phenomenon was observed frequently and elevated effluent (chemostat influent) suspended solids concentrations. Downing *et al.* (2013) and Bassin *et al.* (2016) also indicated that high SOLRs potentially enhanced biofilm detachment rates. In this context, Aygun *et al.* (2008) also reported a plateau occurred in biomass production level after SOLR reached 50 g COD m⁻² d⁻¹.

The trend of VSS/TSS ratio in attached biofilm across the HR-MBBR and inoculum range was 0.67 ± 0.06 and 0.71 ± 0.02 mg VSS/mg TSS, respectively (Table 2). The VSS/TSS ratio obtained by De Oliveira *et al.* (2014) based on pilot scale average values was 0.69 mg VSS/mg TSS; however, this value reported by Jähren *et al.* (2002) operated at lab scale, equal to 0.91 , was much higher. This may be due to the fibrous materials with low VSS/TSS ratio (almost 0.55 mg VSS/mg TSS) carried by raw wastewater and despite biomass adhering to the carriers.

HR-MBBR and IC effluent

The f_{VT} (VSS/TSS) and f_{CV} (X_{COD}/VSS) ratios in the effluent of HR-MBBR and IC processes are shown in Figure 7. The f_{VT} value in the HR-MBBR increased from 0.5 to 0.8 g VSS/g TSS with an increasing SOLR from 2 to 16 g CS_B m⁻² d⁻¹. In the IC process, the f_{VT} value only increased from 0.60 to 0.7 g VSS/g TSS as the SOLR increased from 20 to 90 g CS_B

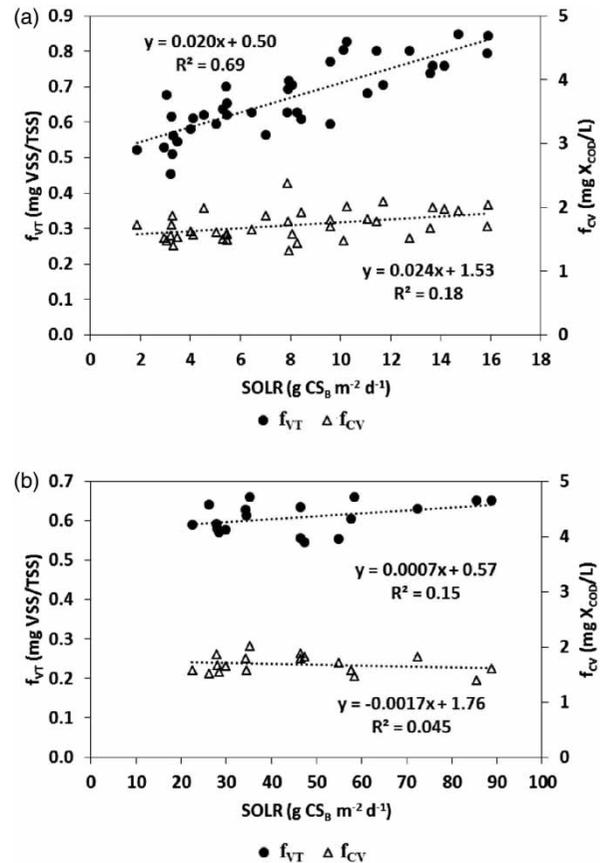


Figure 7 | Effect of SOLR on f_{VT} and f_{CV} ratio for (a) HR-MBBR and (b) IC process effluents.

m⁻² d⁻¹ despite some fluctuations that may have resulted from detached biofilm. The values of f_{CV} in the effluent in both processes' effluent was 1.7 ± 0.2 g X_{COD}/g for all OC.

The f_{VT} value in the effluent reported by Brosseau *et al.* (2016) was 0.81 to 0.91 g VSS/g TSS, while f_{CV} varied between 1.24 and 1.6 g X_{COD}/g for all OC based on HRT and SOLR; the same ratio was also reported by Karizmeh (2012).

The effect of the HRT on the effluent COD fractions was evaluated for the HR-MBBR and IC processes (Figure 8). The particulate matter fraction increased after the HR-MBBR or the IC processes, with the largest proportion observed with an HRT of 36 min in the HR-MBBR and of 225 min in the IC process. The fractionation of COD showed that particulate and soluble COD were predominant in the influent and effluent of HR-MBBR and IC processes, whereas the COD contained a small portion of colloidal matter.

Particle agglomeration occurred with increasing HRT up to 36 and 225 min across the HR-MBBR and IC

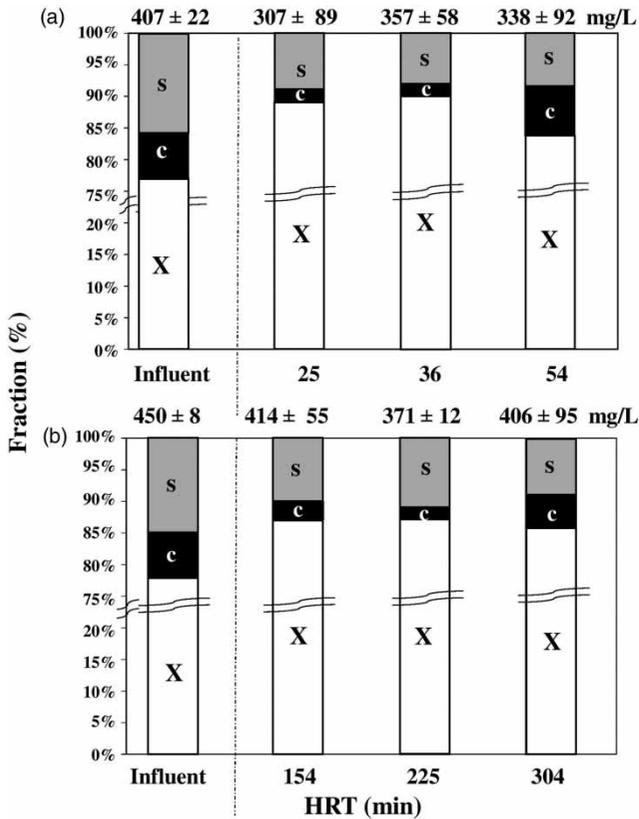


Figure 8 | Influent and effluent mean COD fractions as a function of HRT for the (a) HR-MBBR and (b) IC processes.

processes; however, with increasing HRT from 36 min to 54 min in HR-MBBR, a movement from particulate toward smaller particle (colloidal) matter was observed. The same characteristics in the MBBR effluent were observed by Brosseau *et al.* (2016) between 37 and 40 min HRT.

Particle agglomeration also resulted from increasing the HRT from 0.75 to 4 hours (Melin *et al.* 2002; Åhl *et al.* 2006; Ødegaard *et al.* 2010; Karizmeh 2012), but Karizmeh *et al.* (2014) later demonstrated that by independently decreasing HRT and SOLR, a shift toward smaller particle size was observed. Moreover, during degradation of particulate matter and formation of smaller particles, more surface area of substrate is available for hydrolysis (Dimock & Morgenroth 2006).

The average value of SVI on different operating condition was measured to evaluate the sludge settleability of each process.

Slightly better settling sludge was obtained in the IC process (SVI of 70 ± 11 mL/g) than in the HR-MBBR

(94 ± 10 mL/g). Better flocculating solids may have resulted from the IC process on which configuration most favors the proper maintenance of SVI in a higher SOLR even at lower SRT (SRT_{IC} : 0.6 ± 0.1 d and $SRT_{HR-MBBR}$: 1.5 ± 0.1 d). These results are supported by Liu *et al.* (2004) who demonstrated that low organic loading rate resulted in irregular shape with poor settling characteristics and high SVI value. According to the theory, low substrate concentrations favor the growth of filamentous over floc-forming bacteria (Chudoba *et al.* 1973; Chudoba 1985).

Aeration requirements

The maximum efficiency of HR-MBBR and IC processes bio-transformation were $90 \pm 3\%$ and $77 \pm 3\%$, respectively, which corresponded to HRTs of 36 min and 3.7 hours, SRTs of 1.5 ± 0.1 d and 0.6 ± 0.1 d, and SOLR of 2.0 to 5.5 g CS_B $m^{-2} d^{-1}$ and 22 to 30 g COD $m^{-2} d^{-1}$.

The OUR in the HR-MBBR (50 mg O_2 $L^{-1} h^{-1}$) was determined to be two and a half times greater than in the IC process 20 mg O_2 $L^{-1} h^{-1}$ (accounting for the inoculum process OUR of 10 mg O_2 $L^{-1} h^{-1}$). The total oxygen demand was calculated to be 0.54 ± 0.03 and 0.81 ± 0.07 kg O_2 /kg CS_B added for the HR-MBBR and IC processes, respectively. The blower provided 16 ± 1 m^3/h and 3.5 ± 0.2 m^3/h of air in the HR-MBBR and IC reactors to maintain DO level 3–4 mg O_2/L and 6–7 mg O_2/L , respectively. It should be noted that oxygen transfer rates and efficiency (OTE) at full scale may differ due to the shallow depth of the pilot reactors.

In a high-rate AS process, Jimenez *et al.* (2015) reported that the maximum removal efficiency was obtained at an HRT ranging between 30 and 45 min, at an SRT of 0.6 ± 0.1 d. They observed that the optimal removal of S_B (80%) required an oxygen concentration of 0.38 ± 0.12 kg O_2 /kg COD (based on the use of net oxygen consumption of bio-transformation).

Poor oxygen transfer efficiencies (rapidly rising bubbles) related to the coarse diffuser in the HR-MBBR can also lead to less DO and excessive power requirement compared with the fine diffuser in the chemostat.

The DO concentration dropped below 2 mg/L in the HR-MBBR process within 5 min of the non-aeration period, while this value took almost 30 min in the

chemostat. In this context, much of the energy can be saved across the IC process with DO control strategy by using programmable logic controllers (PLC) for multi loop controllers of the aeration system (turn automatic switching range on/off DO transmitters); however, cost-effectiveness analysis needs to be conducted to provide further results to fully compare the IC and HR-MBBR processes.

From an energy efficiency point of view, operating the IC process as an interesting alternative to high-rate system may lead to diminishing the consumption of energy through the aeration system and also resulted in the efficient production of energy across the anaerobic digester by minimizing hydrolysis of X_B .

CONCLUSION

The objective of this study was to determine the potential of an innovative high-rate IC process compared with a typical HR-MBBR process for colloidal and soluble organic matter transformation into particulate matter for anaerobic digester methane production. The effect of SOLR, OUR, and HRT on the removal and bio-transformation of C_B and S_B fractions was studied using real wastewater in a pilot scale system operated under five OC. The SOLR in the HR-MBBR and IC processes were varied between 2 to $16 \text{ g m}^{-2} \text{ d}^{-1}$ and 20 to $90 \text{ g m}^{-2} \text{ d}^{-1}$ respectively, using different HRTs.

The following conclusions were drawn:

- CS_B bio-transformation into X_B in the HR-MBBR process increased with HRT (and SOLR) up to 36 min with a CS_B capture efficiency as high as $90 \pm 3\%$, while in the IC process, an HRT of 3.7 hours was required for a CS_B capture of up to $77 \pm 3\%$ at SOLRs between 22 and $30 \text{ g } CS_B \text{ m}^{-2} \text{ d}^{-1}$.
- The SOUR value across the HR-MBBR and chemostat, to maintain a DO level above $2 \text{ mg O}_2/\text{L}$, was similar in both systems (55 ± 1 and $53 \pm 6 \text{ mg O}_2 \text{ g}^{-1} \text{ VSS h}^{-1}$, respectively).
- A slightly better settleability of produced particulate matter, based on SVI values, was obtained in the IC process ($70 \pm 11 \text{ mL/g}$) than in the HR-MBBR ($94 \pm 10 \text{ mL/g}$), possibly due to the fact that better flocculating solids may

have resulted from higher SOLR and lower SRT values in the IC than the HR-MBBR process.

The innovative IC process can be a competitive alternative process to maximize the bio-transformation of CS_B to minimize X_B and X_{OHO} oxidation to improve the energy balance at WRRFs.

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