One-step approach for the synthesis of CoFe$_2$O$_4$@rGO core-shell nanocomposites as efficient adsorbent for removal of organic pollutants

Z. J. Song, W. Ran and F. Y. Wei

ABSTRACT

CoFe$_2$O$_4$-reduced graphene oxide nanocomposites (CFG) have been successfully synthesized via one-step solvothermal method. The prepared CFG are characterized by X-ray diffraction, Raman spectroscopy, Fourier transform infrared spectroscopy, field emission scanning electron microscopy (FESEM), vibrating sample magnetometer and so on. The FESEM results show that CFG have uniform core-shell structure with an average diameter of about 75 nm and the thickness of the outer graphene shell is about 15–20 nm. The mass ratio of CoFe$_2$O$_4$ to graphene oxide is a key factor affecting the formation of core-shell hybrids. CFG display much higher adsorption capacity for anionic dyes than cationic dyes owing to the favorable electrostatic interaction. The adsorption capacity for methyl orange is observed as high as 263 mg g$^{-1}$ at 298 K, and the adsorption isotherms follow the Langmuir model. Furthermore, the specific saturation magnetization (Ms) of CFG is 32.8 emu g$^{-1}$, and the as-synthesized nanocomposites can be easily separated by external magnetic field after adsorption. The results suggest that CFG have great potential for the practical industrial wastewater treatment.

Key words | anionic dyes, CoFe$_2$O$_4$-reduced graphene oxide nanocomposites, core-shell structure, one-step method, selective adsorption

INTRODUCTION

In recent years, graphene/magnetic oxide (GM) nanocomposites have drawn tremendous scientific interest for their distinguished properties such as strong superparamagnetism, large surface area and excellent extraction ability. GM composites have been successfully applied in drug carriers (Yang et al. 2009), catalyst (Fu et al. 2012; Yao et al. 2013), lithium storage (Yang et al. 2010; Sun et al. 2014) and other fields (Liu et al. 2013; Xue et al. 2014). In addition, GM composites are excellent adsorbents in the field of water treatment. Until now, the adsorptive removal of dyes and heavy metal ions by this type of nanocomposites has been demonstrated. Li et al. (2011) synthesized magnetic CoFe$_2$O$_4$-functionalized graphene sheets (FGS) nanocomposites by hydrothermal treatment, and discovered that the cobalt ferrite nanoparticles with diameters of 10–40 nm were uniformly distributed on FGS, and the adsorption capacity for removing methyl orange (MO) is 71.54 mg g$^{-1}$. Qi et al. (2015) found that magnetite/reduced graphene oxide (MRGO) nanocomposites obtained by different methods have different characteristics. The MRGO prepared by the co-precipitation method showed special adsorption ability to negative ions, but those synthesized by the solvothermal method had the best extraction ability and reusability to metal ions. Nevertheless, these reported magnetic nanoparticles are loaded on the surface of graphene sheets (Xie et al. 2012; Li et al. 2014), and may have the defects of detachment and aggregation of magnetic metal oxides from graphene sheets, which decreased their saturation magnetization (Fu et al. 2015; Zhan et al. 2015), and reduced their efficacy in practical applications. One of the most promising strategies to tackle the aggregation problem of metal oxides is to enwrap them with a polymer or silica shell (Dong et al. 2006; Luo et al. 2009). However, the reported materials enwrapped with the polymer or silica shell showed low saturation magnetization. Therefore, the fabrication of core-shell magnetic nanocomposites with both strong magnetic responsivity and high absorption capacity remains challenging. Core-shell structure Fe$_3$O$_4$@graphene oxide (GO) submicron particles have

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been prepared via the two-step electrostatic self-assembly process and exhibited large adsorption capacity for bovine serum albumin (BSA) (Wei et al. 2012).

In this paper, we develop a simple solvothermal route to prepare the core-shell structure CoFe2O4@rGO (abbreviated as CFG) nanocomposites, in which the reduction of GO and the crystallization of CoFe2O4 crystals happened in one step without adding any reducing agent. The formation mechanism of the core-shell structure was initially discussed. The adsorption properties were also measured using MO as model pollutant in aqueous solutions. The core-shell nanocomposites can suppress the aggregation of oxide nanoparticles, maintain a high saturation magnetization and give rise to a high adsorption capacity. The obtained CFG can be recyclable, and thus has great potential for the removal of toxic pollutants from wastewater.

**EXPERIMENTS**

**Materials**

Graphite powder, ethylene glycol (EG), polyethylene glycol 2000 (PEG 2000), anhydrous ferric chloride (FeCl3), cobalt (II) chloride hexahydrate (CoCl2·6H2O), methylene blue (MB), MO, Congo red (CR), Rhodamine B (RhB) and other reagents were purchased from Sinopharm Chemical Reagent Co. (Hefei, China). All chemicals used here are analytical grade without further purification.

**Preparation of CFG**

GO was synthesized by the modified Hummers method (Marcano et al. 2010). CFG were prepared by solvothermal method with minor modifications (Xue et al. 2014). In a typical synthesis, the as-prepared GO (0.291 g) was exfoliated by ultrasonication in 70 mL of EG and 2.5 g of polyethylene glycol 2000 (PEG 2000) mixture at 75 °C, followed by the addition of FeCl3 (2 g) and CoCl2·6H2O (1.42 g) to form a brown mixture solution. 30 mL of EG solution containing 9 g of NaAc and 1 g of NaOH were slowly added into the above brown solution, followed by vigorous stirring for 2 h. The whole process was kept ultrasonic to avoid the aggregation of the GO in EG. After that, the resulting mixture was sealed in a Teflon-lined stainless-steel autoclave, maintained at 453 K for 24 h, then cooled naturally to room temperature. The black products were obtained by magnetic separation, washed several times with absolute ethanol and deionized water and dried at 333 K for 12 h.

**Characterization**

The powder X-ray diffraction (XRD) patterns of the samples were performed using a Rigaku D/max 2500 V X-ray diffractometer with CuKα radiation (λ = 1.5418 Å). The morphology and structure of the samples were characterized by a field emission scanning electron microscope (FESEM, Hitachi SU8020) and high-resolution transmission electron microscope (HRTEM, JEM-2100F). Fourier transform infrared spectroscopy (FTIR) spectra were recorded on a Perkin-Elmer Spectrum 100. X-ray photoelectron spectroscopy (XPS) was conducted using an ESCALAB 250 spectrometer. The magnetic properties of samples were studied using a vibrating sample magnetometer (VSM, PPMS-14T) at a temperature of 300 K. The zeta potential was measured by the ZETASIZER Nano-ZS from Malvern Instruments.

**Adsorption experiments**

The adsorption of dyes in aqueous solution on CFG was performed in a batch experiment. Typically, 0.025 g of the adsorbent was added into 25 mL of MO solutions of the desired initial concentrations (10 mg L⁻¹) and then the mixture was shaken at a speed of 180 rpm at 298 K for 2 h. The nanocomposites were removed from the solution by magnetic separation, and the concentrations of MO, CR, MB, and RhB were determined by a Vis spectrophotometer (722, China) at 465, 497, 662 and 550 nm, respectively. The adsorption capacity qe and removal efficiency R were calculated by the following expressions:

\[
q_e = \frac{(C_0 - C_e)V}{m} \\
R = \frac{(C_0 - C_e)}{C_0} \times 100\% 
\]

where \(q_e\) (mg g⁻¹) represents the adsorption capacity at equilibrium, \(C_0\) and \(C_e\) (mg L⁻¹) are the initial and the equilibrium concentration of the dye remaining in the solution. \(V\) (L) is the volume of the aqueous solution, and \(m\) (g) is the dry weight of composites.

**Reusability experiment**

CFG (0.1 g) was added to 50 mL of MO solution (100 mg L⁻¹) and then the mixture was shaken for 1.5 h at room temperature. After separation of the nanocomposites by an external...
magnet, the supernatant solution was analyzed. The nanocomposites’ adsorbent with MO was washed with absolute ethanol solution several times, collected by a magnet and reused for the next adsorption experiment. The reusability experiments were performed six times.

RESULTS AND DISCUSSION

Structure and morphology of samples

The phase structures of as-synthesized samples were firstly determined by XRD. As shown in Figure 1(a), the GO shows a sharp peak at $2\theta = 9.4^\circ$ corresponding to the (001) reflection of GO. By contrast, the characteristic peak of GO cannot be detected in the XRD pattern of the CFG (Figure 1(b)), suggesting that GO was effectively reduced (Yao et al. 2015b). Meanwhile, the main diffraction peaks at $2\theta = 30.1, 35.3, 42.7, 56.8$ and $62.2^\circ$ match well with the cubic spinel structure of cobalt ferrite (JCPDS card no. 22-1086), and correspond to crystal indexes of (220), (311), (400), (511), and (440), respectively (Wan & Jian 2015).

Figure 2 shows a comparison of Raman spectra of CoFe$_2$O$_4$ and CFG, recorded in the frequency range 100–2,200 cm$^{-1}$. For all the samples, three major Raman active modes of broadening peaks are detected clearly at about 298, 456, and 666 cm$^{-1}$, which shows that the prepared nanocomposites consisted of CoFe$_2$O$_4$ (Ding et al. 2015). In comparison with the CoFe$_2$O$_4$, CFG shows a Raman shift around 1,352 and 1,593 cm$^{-1}$ corresponding to the D band and G band of grapheme oxide (Gan & Shang 2015). The existence of these two bands in CFG shows that the graphene has been incorporated into the nanoparticle system (An et al. 2014). Meanwhile, the ratio of intensities ($I_D/I_G$) represents the structure defect degree of the graphite materials. The value of $I_D/I_G$ for GO was 0.93. For CFG, that was markedly increased to 1.06, indicating formation of some sp$^3$ carbon by functionalization as a large number of oxygen groups were removed in the solvothermal reduction process (Joon et al. 2012).

FTIR spectra of GO, CoFe$_2$O$_4$, and CFG are shown in Figure 3. Several characteristic peaks of GO can be observed in Figure 3(a), confirming the successful oxidation of graphite. In detail, the intense band at 3,423 cm$^{-1}$ originated from O–H stretching vibration. The bands at 1,735 and 1,620 cm$^{-1}$ are associated with C=O stretching of COOH groups located at edges of GO sheets. The bands at 1,196 and 1,050 cm$^{-1}$ are attributed to the stretching of epoxy C–O and alkoxy C–O band (Xie et al. 2012), respectively. By comparison, most bands related to the oxidized groups vanish in the FTIR (Figure 3(c)) of CFG except for two weak peaks at 3,423 cm$^{-1}$ (v (O–H)) and 1,082 cm$^{-1}$
revealing that GO was effectively reduced in the synthesis process. For CoFe$_2$O$_4$ (Figure 3(b)), the peak at around 598 cm$^{-1}$ was observed, which is assigned to the Fe(Co)–O band in the cobalt ferrite (Lu et al. 2014). The absorption peak at 1,640 cm$^{-1}$ is ascribed to the vibration of adsorbed water (Zhang et al. 2015). The new adsorption band at 1,567 cm$^{-1}$ in FTIR of CFG can be assigned to the stretching vibrations of the un-oxidized carbon backbone (Ji et al. 2011). Meanwhile, the stretching vibration of Fe(Co)–O band at 598 cm$^{-1}$ gets red shift to 582 cm$^{-1}$.

To investigate the morphology and particle size of the products, Figure 4 gives the corresponding FESEM and HRTEM images. The FESEM image of GO (Figure 4(a)) displays a wrinkled and scrolled morphology paper-like structure of the ultrathin graphene sheets and stacking of sheets. For the CoFe$_2$O$_4$ nanoparticles (Figure 4(b)), uniform spheres with an average diameter of approximately 120 nm are observed. It is found that their diameters are mainly distributed in the range of 80–160 nm (Figure 5(a)). From Figure 4(c), CFG becomes smaller with an average diameter of about 75 nm, the diameter distribution being in the range of 50–105 nm (Figure 5(b)). Furthermore, compared with CoFe$_2$O$_4$ nanoparticles, CFG becomes more irregular. It may contribute to retarding the crystalline rate of CoFe$_2$O$_4$ by the graphene sheets during the solvothermal process. The HRTEM image of the CFG (Figure 4(d))

![Figure 3](https://iwaponline.com/wst/article-pdf/75/2/397/456040/wst075020397.pdf)

**Figure 3** | FTIR spectra of (a) GO, (b) CoFe$_2$O$_4$, and (c) CFG.

![Figure 4](https://iwaponline.com/wst/article-pdf/75/2/397/456040/wst075020397.pdf)

**Figure 4** | FESEM images of (a) GO, (b) CoFe$_2$O$_4$, (c) CFG, and (d) HRTEM image of CFG.
indicates that the CoFe$_2$O$_4$ nanoparticles are coated with a uniform shell of graphene with the thickness of approximately 15–20 nm and the diameter of the CoFe$_2$O$_4$ core is around 50 nm. This means that a novel type of core-shell structure CFG has been successfully synthesized.

In the present work, it is worth noting that the magnetic CoFe$_2$O$_4$ nanoparticles decorated onto graphene sheets are obtained under the same conditions only by decreasing the mass ratio of CoFe$_2$O$_4$ to GO to 3. According to previous reports (Sharifi et al. 2015), there are several factors that may contribute to the formation of the core-shell CFG. First, the existence of the van der Waals forces between the adjacent magnetic CoFe$_2$O$_4$ nanoparticles make them move towards each other and GO sheets are aggregated simultaneously. Second, the graphene sheet can roll itself up into a scroll structure under ultrasonic effect. Third, the nanoparticles attached at the edge of the GO sheets may be attracted by the inner part of the GO sheets due to sparsely distributed oxygen-containing functional groups preceding the rolling (Xue et al. 2014). Although the detailed mechanism of the growth of graphene-CoFe$_2$O$_4$ core-shell nanocomposites is not very clear, it is obvious that the mass ratio of CoFe$_2$O$_4$ to GO is a key factor for the formation of core-shell hybrids.

**Vibrating sample magnetometer analysis**

To evaluate the magnetic separation capacity of CFG, we conducted the magnetic hysteresis loops of CFG and pure CoFe$_2$O$_4$ at 300 K, as shown in Figure 6. The saturation magnetization (Ms) of the pure CoFe$_2$O$_4$ nanoparticles is 45.4 emu g$^{-1}$, whereas the Ms value of CFG is 32.8 emu g$^{-1}$, which is slightly lower than that of pure CoFe$_2$O$_4$, but much higher than that of the magnetic metal oxides-graphene sheets hybrid composites (Li et al. 2011; Xie et al. 2012). The HRTEM image shows that CFG possesses a magnetically inactive shell of graphene, thus affecting the magnitude of magnetization. The magnetic separability of CFG was also tested by placing a magnet near the glass bottle. The black product was rapidly attracted toward the magnet in a short period (inset in Figure 6), demonstrating high magnetic sensitivity of CFG.

**Removal of dyes from aqueous solution**

The adsorption performances of four organic dyes (two cationic dyes and two anionic dyes) on CFG were also studied.
The adsorption isotherms at room temperature are shown in Figure 7. The data of maximum adsorption for these dyes are summarized in Table 1. The results show that CFG have stronger affinity for anionic dyes than for cationic dyes.

Zeta potential of adsorbent is a key factor to influence its adsorption capacity, and thus is measured to understand further why CFG can remove anionic dye more effectively than cationic dye. The zeta potential value of CFG synthesized using EG as solvent is about 30.3 mV, indicating that the surface of CFG is positively charged. The high absolute zeta potential value also proves that the adsorbent will be stable in aqueous solution (Fan et al. 2013; Qi et al. 2018). In addition, we find that the zeta potential value of CFG synthesized with EG/H2O mixed solution as solvent will drop to ~25.3 mV, and thus the adsorption capacity for anionic dyes is obviously decreased.

Based on the above, the mechanism of the adsorption of organic dyes by CFG is probably due to electrostatic interaction. For instance, MO is an anionic azo dye which contains the sulfonic acid groups (R-SO3Na). In aqueous solution, dye dissociates to the sulfonate anions (R-SO3-) and the sodium ions (Na+). Therefore, R-SO3- ions were easily absorbed to the surface of CFG by electrostatic force. Although the detailed formation mechanism of CFG surface charge was not very clear, it is certain that the type of reaction media caused some surface modifications onto CFG owing to the difference in the surface tension and dielectric constant (Ramesha et al. 2011; Hayyan et al. 2015).

Desorption and reuse of adsorbents

The recycling and regeneration ability is significant for the practical application of adsorbents. Therefore, the recycling of CFG in the removal of MO was investigated. Figure 8 shows the removal efficiency of MO on CFG as a function of cycling runs. The values of R after 1.5 h are 98.5, 96.1, 95.8, 92.2, 90.3 and 88.6% for the consecutive six cycles, respectively. The decline of removal efficiencies is not

![Figure 7](image1.png)

**Figure 7** Adsorption isotherms of dyes on CFG.

![Figure 8](image2.png)

**Figure 8** The relationship of the removal efficiency of MO on CFG with cycles.

![Figure 9](image3.png)

**Figure 9** Adsorption isotherms of MO on CFG.

<table>
<thead>
<tr>
<th>Dye</th>
<th>qmax (mg·g⁻¹)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Anionic</td>
<td>MO 246, CR 491</td>
</tr>
<tr>
<td>Cationic</td>
<td>MB 38, RhB 21.5</td>
</tr>
</tbody>
</table>

![Table 1](image4.png)

**Table 1** The maximum adsorption capacities qmax of different dyes on CFG
more than 10% after six recycles, indicating that CFG have a good reusability.

**Adsorption isotherms analysis of MO**

The adsorption isotherms of MO on CFG are shown in Figure 9 at 298, 308 and 318 K. The Langmuir isotherm model (Yao et al. 2012a, 2012b) is used to describe the monolayer adsorption process, and the model can be described in linear form by the following equation:

\[
\frac{c_e}{q_e} = \frac{1}{q_{\text{max}}b} + \frac{1}{q_{\text{max}}}c_e
\]

where \(q_{\text{max}}\) (mg g\(^{-1}\)) is the maximum adsorption capacity at complete monolayer coverage, \(b\) (L mg\(^{-1}\)) is the Langmuir adsorption constant.

The Freundlich isotherm model (Sharifi et al. 2013) shows that the multilayer of the adsorption process occurs on heterogeneous surfaces, and is expressed as follows:

\[
\ln q_e = \ln K_F + \frac{1}{n} \ln c_e
\]

where \(K_F\) (mg\(^{1-n}\)L\(^n\)g\(^{-1}\)) represents the adsorption capacity, and \(n\) represents the degree of dependence of adsorption with equilibrium concentration.

The related parameters of two models are calculated and displayed in Table 2. From the linear correlation coefficients \(R^2\) at different temperatures, it can be found that the Langmuir model is more suitable for describing the adsorption of MO on CFG compared to the Freundlich model. In addition, the F values obtained from the Langmuir equation are far higher than those from the Freundlich equation. The results suggest that monolayer adsorption of MO on CFG is the main mechanism (Li et al. 2012). Besides, the values of \(q_{\text{max}}\) calculated from the Langmuir model are the highest at \(T = 298\) K and the lowest at \(T = 318\) K, which indicates that increasing temperature is unfavorable to the adsorption and the adsorption process is exothermic.

Calculated from the Langmuir isotherm model, the maximum equilibrium adsorption capacity \(q_{\text{max}}\) of MO on CFG is 263 mg g\(^{-1}\) at 298 K. A comparison of \(q_{\text{max}}\) for MO uptake between CFG and different magnetic adsorbents reported in the literature can be seen in Table 3. It is obvious that CFG synthesized in this work exhibits quite a high \(q_{\text{max}}\), which suggests that the as-synthesized CFG can be considered as an effective adsorbent for the removal of MO dye from wastewater.

**Adsorption thermodynamics**

Thermodynamics parameters including Gibbs free energy change \(\Delta G\), enthalpy change \(\Delta H\), and entropy change \(\Delta S\) are calculated from the adsorption isotherms (Figure 9). \(\Delta G\) is calculated using Equation (5). \(\Delta H\) and \(\Delta S\) are determined from the slope and intercept of the plot of \(\ln K_0\) versus \(1/T\) using Equation (6):

\[
\Delta G = -RT \ln K_0
\]

### Table 2 | Parameters for Langmuir and Freundlich isotherm models

<table>
<thead>
<tr>
<th>(T (K))</th>
<th>(q_{\text{max}}) (mg g(^{-1}))</th>
<th>(b) (L mg(^{-1}))</th>
<th>(R^2)</th>
<th>(F)</th>
</tr>
</thead>
<tbody>
<tr>
<td>298</td>
<td>263</td>
<td>0.09</td>
<td>0.9996</td>
<td>17,374</td>
</tr>
<tr>
<td>308</td>
<td>246</td>
<td>0.07</td>
<td>0.99873</td>
<td>5,515</td>
</tr>
<tr>
<td>318</td>
<td>220</td>
<td>0.04</td>
<td>0.99865</td>
<td>5,181</td>
</tr>
</tbody>
</table>

### Table 3 | Summary of MO maximum adsorption capacities on various magnetic adsorbents

<table>
<thead>
<tr>
<th>Adsorbent</th>
<th>(q_{\text{max}}) (mg g(^{-1}))</th>
<th>(T/K)</th>
<th>Reference</th>
</tr>
</thead>
<tbody>
<tr>
<td>ZnLa(<em>{0.02})Fe(</em>{1.98})O(_4/)MWCNTs</td>
<td>81</td>
<td>298</td>
<td>Zhang &amp; Nan (2015)</td>
</tr>
<tr>
<td>Co/MWCNTs</td>
<td>170</td>
<td>293</td>
<td>Zhao et al. (2015)</td>
</tr>
<tr>
<td>MPGM</td>
<td>80</td>
<td>298</td>
<td>Wang et al. (2015)</td>
</tr>
<tr>
<td>MWCNT/SPIONs</td>
<td>10.89</td>
<td>298</td>
<td>Bayazit (2014)</td>
</tr>
<tr>
<td>CoFe(_2)O(_4/)FGS</td>
<td>71.54</td>
<td>300</td>
<td>Li et al. (2011)</td>
</tr>
<tr>
<td>m-CS/g-Fe(_2)O(_3/)MWCNTs</td>
<td>66</td>
<td>293</td>
<td>Shuang et al. (2012)</td>
</tr>
<tr>
<td>CFG</td>
<td>263</td>
<td>298</td>
<td>This work</td>
</tr>
</tbody>
</table>
\[ \ln K_0 = \frac{\Delta S}{R} - \frac{\Delta H}{RT} \]

where \( R \) and \( T \) represent the universal gas constant \((8.314 \text{ J mol}^{-1} \text{ K}^{-1})\) and the system temperature (K). \( K_0 \) is the equilibrium distribution coefficient that is obtained by plotting \( \ln(q_e/C_0) \) versus \( q_e \) and extrapolating \( q_e \) to zero.

The thermodynamic data calculated by Equations (5) and (6) are shown in Table 4. The value of \( \Delta H \) during the adsorption process is calculated to be \(-34.8 \text{ kJ mol}^{-1}\), which indicates that the adsorption process is exothermic. Meanwhile, the \( \Delta S \) values are all positive at different temperatures, which illustrate the increasing randomness at the solid-solution interface during fixation of MO onto the surface of CFG. The negative values of \( \Delta G \) demonstrate a spontaneous and feasible process for the adsorption. The value of \( \Delta G \) becomes more negative with the decrease of temperature, which indicates that the adsorption is suitable at low temperature.

### CONCLUSIONS

In summary, we have synthesized a novel type of core-shell structure CoFe\(_2\)O\(_4@r\)GO nanocomposites (CFG) via a facile one-step solvothermal route. The as-synthesized CFG have a uniform core-shell structure with an average diameter of about 75 nm, where the CoFe\(_2\)O\(_4\) nanoparticles are coated by a layer shell of GO with a thickness of approximately 15–20 nm. The mass ratio of CoFe\(_2\)O\(_4\) to GO is a key factor for the formation of core-shell hybrids. CFG displays much higher adsorption capacity for anionic dyes owing to electrostatic interaction between CFG and dyes. Compared with previous literature, CFG exhibits a very high absorption capacity of 263 mg g\(^{-1}\) for MO at 298 K, and the prepared CFG nanocomposites can be easily separated by an external magnetic field after adsorption. After six cycles of the adsorption–desorption process, their adsorption capacity decreased slightly. This work indicates that CFG have great potential for organic pollutants wastewater treatment.

### ACKNOWLEDGEMENTS

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